

Predicting intervention priorities for wildlife conflicts

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Abstract:	<p>There is growing interest in developing effective interventions to manage socially- and environmentally-damaging conservation conflicts. Recent studies have identified a wide variety of different intervention strategies in various contexts but the reasons why one type of intervention is chosen over another remain underexplored. In this international study we surveyed conservation researchers and practitioners (N=427) to explore how the characteristics of conflicts and characteristics of decision-makers influence conflict recommendations. Using a fully-factorial design, we experimentally manipulated three aspects of eight different conflict scenarios – the development status of the country, the conflict framing, and whether wildlife killing was illegal – and recorded whether respondents prioritised one of five intervention types: wildlife impact reduction, awareness, enforcement, economic incentives or stakeholder engagement. We also recorded information on respondents' demographic and disciplinary backgrounds. Stakeholder-based interventions were recommended most often in the survey and in written feedback. However, fitting multinomial mixed logit models with no missing scenarios (N=411), we find that recommendations are influenced by small changes in the details of conflict, and differ according to respondent characteristics. Enforcement and awareness interventions are prioritised more in conflicts in more highly developed nations and by respondents with more natural-science backgrounds and less experience of conflicts. Contrastingly, economic interventions are prioritised more when wildlife killing is described as illegal. Respondent age, gender and the development status of their home country also predicted some intervention decisions. Further interrogating the influences shaping conservation decision-making will help towards developing evidence-informed interventions.</p>

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3

4 **Abstract**

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24 conservation decision-making will help towards developing evidence-informed interventions.

25 Introduction

26 Conservation conflicts are damaging for both people and wildlife and as such, there is much interest
27 in designing and implementing interventions to resolve or mitigate them (Redpath et al. 2013).

28 Although conservation conflicts can involve clashes over any conservation objective (Redpath et al.
29 2015) conflicts centring on the impacts of wildlife on livelihoods are particularly widespread (Pooley
30 et al. 2016). In these situations – which are often framed as ‘human-wildlife conflict’ or ‘coexistence’
31 problems – interventions commonly aim to mitigate the negative impacts of wildlife, reduce wildlife
32 killings or improve the relationships between stakeholders (Baynham-Herd, et al. 2018).

33
34 Recent research has explored the geographical distribution of interventions (Ravenelle & Nyhus
35 2017) and assessed their effectiveness (van Eeden et al. 2018; Eklund et al. 2017). Other studies
36 have identified variation in how practitioners and researchers prioritise different interventions
37 (Rastogi et al. 2013; Shiffman & Hammerschlag 2016). For instance, how conflicts are framed by
38 authors, whether they involve illegal behaviours and the development status of the countries in
39 which they are located have been hypothesized as to influence intervention decisions (Baynham-
40 Herd, et al. 2018; Soliku & Schraml 2018). Moreover, it appears that researchers and practitioners
41 from different disciplinary backgrounds and regions tend to recommend different solutions (Lute et
42 al. 2018). However, the underlying reasons accounting for this variation in intervention priorities has
43 been less explored, in part because much previous work in this area has been observational, making
44 it harder to unpick potential relationships. Moreover, as intervention strategies used in conflicts can
45 often be contested or controversial (López-Bao et al. 2017; Duffy et al. 2019), it is important to
46 understand the factors driving support for such different approaches.

47

48 One pathway to better understanding how decisions are made in conflicts is through exploring the
49 social and psychological mechanisms underpinning conservation decision making (Papworth 2017).
50 For instance, subtle changes in the way problems are framed often change how people suggest
51 solving them (Sapiains et al. 2016). Such subtleties may be particularly important when people are
52 making quick decisions with limited information. This is because under such circumstances people
53 are thought to rely more on intuition and pattern matching compared to when making slower, more
54 analytical decisions, using multiple sources of information (Evans 2008; Kahneman 2011).
55 Furthermore, it is known, that like all people (Schultz 2011), the priorities of conservation
56 professionals differ (Sandbrook et al. 2019) and these are likely shaped by predispositions, cognitive
57 biases and values (Sheil & Meijaard 2010; Kiik 2018). However, how such factors might influence
58 conflict intervention decision-making remains underexplored.

59

60 The purpose of this study is to test how particular characteristics of conflicts and of decision-makers
61 influence conflict intervention priorities. To do this we conducted an experimental survey with
62 conservation researchers and practitioners internationally (N=427), in which we presented
63 participants with eight different conflict scenarios, and asked them to prioritise one (out of five)
64 intervention types to manage the conflict in each scenario. Drawing upon similar strategies used in
65 choice (Keane et al. 2016) and framing experiments (Sapiains et al. 2016), by offering limited
66 information per scenario we aimed to identify possible predispositions and tacit influences on
67 decisions. Using a fully-factorial design, we experimentally manipulated three factors hypothesized
68 to influence conflict decisions: the framing of the conflict as being between people and wildlife or
69 between groups of people, whether behaviours were reported as illegal, and the development
70 status of the country where the conflict occurs. We experimentally manipulated these three factors
71 – rather than other relevant factors such as taxa or types of impacts – because their potential

72 influence had been highlighted in a previous review (Baynham-Herd, et al. 2018) but had not yet
73 been tested.

74

75 We then used multinomial mixed logit regressions to test whether these manipulations and the
76 characteristics of participants predicted intervention recommendations. These related to
77 respondents' disciplinary and professional background, and experience with conflicts on the ground
78 and in the literature – which we hypothesised might influence how respondent's conceptualised
79 conflicts, and subsequently the extent to which stakeholder-interventions were prioritised in
80 particular. We also recorded respondents' nationality, gender and age to determine the
81 demographic of the sample given that personal characteristics of decision-makers has been found to
82 shape conservation priorities, preferences and outcomes (Keane et al. 2016; Sandbrook et al. 2019).
83 We then interpreted these results in light of qualitative insights derived from respondents' written
84 feedback.

85

86

87 **Methods**

88 **Survey design**

89 We designed and carried out a short (5-10 minute) online survey using the platform 'Qualtrics'
90 (qualtrics.com). We used an online survey, rather than a written survey to allow for greater flexibility
91 over survey design (including randomization of the specific sub-set of scenarios presented to
92 participants), to reduce the risk of biased responding (of socially undesirable answers) (Gnambs &
93 Kaspar 2014) and to enable the survey to be disseminated internationally. The survey included an
94 information sheet, a series of demographic questions, and then it presented participants with eight
95 different conflict scenarios in turn (Supporting Information, Appendix 3). Each scenario related to a

96 real-world conflict described in the literature, involved one particular species of conservation
97 concern, and some kind of human activity that was threatening the species. The number of scenarios
98 was constrained by survey-length, and the cases involved were selected on the basis of: a) appearing
99 in the conflict literature, b) having species ranges that encompassed at least one very highly
100 developed country and one less highly developed country, and c) reflecting a mix of herbivorous and
101 carnivorous, marine and terrestrial mammals and non-mammals (Table 1).

102

103 For each scenario, participants were asked to select one of five different intervention types, which
104 they deemed of highest priority in that scenario. Following Baynham-Herd et al., (2018) we included
105 five different conflict interventions types: wildlife impact reduction, awareness or training programs,
106 enforcement, economic incentives or compensation and stakeholder engagement. Scenarios (<100
107 words) and intervention options (<15 words) were described in brief and consistent manner and
108 appeared in the same order for each participant (Figure 1).

109

110 Between participants, a full-factorial design was used to systematically vary three aspects of
111 scenario descriptions including: whether they were framed as human-human or human-wildlife
112 conflict, whether wildlife killing was described as illegal and the country the conflict was located in
113 (Figure 1). Each scenario was adapted from existing literature and different pairs of countries were
114 chosen on the basis of maximizing the variation in development status (as determined by the Human
115 Development Index (HDI) (UNDP 2016)), whilst keeping within a given species' range (IUCN 2017).
116 The final combination of scenarios was chosen to ensure a geographical spread across world regions
117 (Table 1). For two scenarios – 'geese' and 'vulture' – the precise species was not named as the
118 conflicts in question related to different, but functionally similar species.

119

120 In each survey, we randomly varied the set of questions (A to H) seen by each participant using the
121 question block randomization feature on Qualtrics. We also included questions on characteristics of
122 the participants, including their disciplinary background, career role and position, nationality,
123 gender, age and familiarity with conflicts in the literature and on the ground. Lastly, we included a
124 section for participants to give open-ended written feedback on both the survey design (e.g.,
125 intervention options) and the factors influencing their decisions.

126

127

128 **Participant recruitment**

129 We first conducted a pilot study at the Scottish Conservation Conflict Research Group
130 (<https://www.conservationconflicts.info/>) meeting in June 2018. After adapting the survey design
131 we then recruited research participants at the European Congress for Conservation Biology in
132 Finland, June 2018 – which was attended by international delegates with varying experience in
133 conflicts and backgrounds. To include a wider range of responses, we also conducted a literature
134 search in ISI Web of Knowledge to identify authors who had recently published studies related to
135 conflicts, and emailed each corresponding author (N=335) asking them to complete the survey and
136 invited people to share the survey on Twitter and via relevant mailing lists. Participants were invited
137 to share their email address (to receive results) but participant anonymity was preserved. In total we
138 received 634 responses. For analysis, we omitted those who identified as ‘not working in
139 conservation’ (N=14) and insufficiently completed responses (<97% completed) leaving a sample of
140 427. For our models, we only including responses with all scenarios eight answered (N=411).
141 Participants came from 52 countries (Supporting Information Appendix 1) and from across different
142 career stages and ages (Table 1), with 84 respondents identifying as ‘practitioners’ or ‘other’ and 321
143 as ‘researchers’. This study received ethical approval from the University of Edinburgh School of
144 Geosciences Ethics panel.

145

146

147 **Analysis**

148 We carried out statistical analysis using the statistical programming software 'R' (R Development
149 Core Team 2016) and the package 'mlogit'. To analyse how different predictors influenced the
150 choices between the five intervention categories, we used multinomial logit linear regressions, with
151 random-parameters to model the correlation between multiple responses (N=8) from each
152 individual. We used the stakeholder intervention type as the reference intervention in reported
153 models (Figure 1), but each other intervention type was used as a reference level in other models for
154 comparison (Table 3).

155

156 Due to some missing responses, models with more variables had slightly reduced sample sizes.
157 Explanatory variable collinearity was checked using Spearman's rho for numerical variables and one-
158 way ANOVA's for categorical variables. As 'Age' was associated with both 'Position' ($F_{2,400} = 183.90$, P
159 < 0.01) and 'Gender' ($F_{1,407} = 35.42$, $P < 0.01$) only the numerical variable 'Age' was included in
160 models. As 'Role' was associated with 'Ground Experience' ($F_{1,356} = 7.081$, $P < 0.01$), only the
161 numerical variable 'Ground Experience' was included in the models. 'Gender' was analysed
162 separately in models (Model set 3) without 'Age'.

163

164 We analysed the data collected from open-ended questions using the software package 'NVivo'.
165 Using a directed content analysis approach (Hsieh & Shannon 2005), we first grouped responses
166 according to whether they addressed pre-determined themes (each intervention type, development
167 status, legality, framing and taxa). Next, using an inductive approach, we added new themes and
168 sub-themes encompassing other commonly discussed subject areas which emerged during analysis

169 (e.g., intervention combinations). We then calculated the frequency of respondents whose feedback
170 was recorded in each given category and reflected upon the content of the prevailing themes with
171 regards to our survey results and interpretation.

172

173

174 **Results**

175 Across the analysed sample (411 participants, 3,288 decisions), the stakeholder intervention type
176 was the most popular but most people varied their priorities across scenarios. Stakeholder
177 interventions were chosen 27% of the time, followed by awareness (25%), economic (20%), wildlife
178 impact reductions (19%) and enforcement (9%). We found that 92% of participants chose at least
179 two of the five intervention type and, 85% chose at least three. Of those who did not deviate from
180 one intervention type (N=33), 85% chose stakeholder only, 6% enforcement only, 3% awareness, 3%
181 wildlife impacts, and 3% economic only.

182

183 Intervention priorities varied dramatically across different conflict taxa. We found that 56% of
184 participants recommended awareness interventions in the vulture conflict scenarios, but only 8% of
185 participants did so for the wolf conflicts. Likewise, 49% of respondents suggested economic
186 interventions in the wolf conflicts, compared to 1% for crocodile conflicts. Enforcement was most
187 popular in the manatee conflicts (25%) and least in the geese conflicts (2%). Stakeholder
188 interventions were most popular for sea otter conflicts (39%) and least for wolf conflicts (18%).
189 Impacts-based interventions were favoured most in bear conflicts (49%) and least in the vulture
190 conflicts (4%) However, intervention decisions varied across the two locations in each scenario
191 (Figure 2).

192

193 Intervention prioritisations were predicted by the development status of the conflict location and
194 whether illegal activity was reported, but not by the conflict framing variable (Figure 3). These
195 effects were consisted across multinomial mixed logit regression models which controlled for the
196 multiple responses per individual, respondent's question blocks and the independent effect of each
197 scenario (Model Set 1, N=411), and those that also included the characteristic of respondents
198 (Model Set 2, N=341). Below, for each predictor variable, results are reported in order of decreasing
199 effect size (odds ratio).

200

201 The higher the HDI of the conflict location the more enforcement and awareness were prioritised.
202 With increasing HDI, the likelihood of choosing enforcement increased compared to economic
203 interventions ($p < 0.01$, Odds Ratio 1.43, 0.95 CI: 1.13-1.79), or impacts ($p < 0.01$, Odds Ratio 1.33,
204 0.95 CI: 1.08-1.67), or stakeholder interventions ($p < 0.05$, Odds Ratio 1.31, 0.95 CI: 1.06-1.63).
205 Similarly, the likelihood of choosing awareness increased compared to economic interventions ($p <$
206 0.01 , Odds Ratio 1.26, 0.95 CI: 1.07-1.47), or impacts ($p < 0.05$, Odds Ratio 1.18, 0.95 CI: 1.02-1.36).
207 When wildlife killing was described as illegal, the likelihood of choosing economic interventions
208 increased compared to awareness ($p < 0.01$, Odds Ratio 1.52, 0.95 CI: 1.12-2.08), or impacts ($p <$
209 0.05 , Odds Ratio 1.49, 0.95 CI: 1.07-2.07), or stakeholder ($p < 0.05$, Odds Ratio 1.45, 0.95 CI: 1.05-
210 1.99) (Table 2, Figure 4).

211

212 The characteristics of respondents also predicted intervention priorities. The more respondents'
213 disciplinary backgrounds were weighted towards natural science over social science, the more likely
214 they chose enforcement and awareness. Specifically, discipline most strongly predicted the
215 likelihood of enforcement being chosen compared to stakeholder ($p < 0.01$, Odds Ratio, 1.47, 0.95
216 CI: 1.21-1.78), or economic interventions ($p < 0.01$ Odds Ratio, 1.33 0.95 CI: 1.09-1.64). Similarly,

217 discipline predicted the likelihood of awareness being chosen compared to stakeholder ($p < 0.01$,
218 Odds Ratio, 1.36, 0.95 CI: 1.18-1.56), or economic ($p < 0.01$, Odds Ratio, 1.38, 0.95 CI: 1.18-1.63) or
219 to a lesser extent, impacts ($p < 0.05$, Odds Ratio, 1.21, 0.95 CI: 1.04-1.40).

220

221 As experience of conflicts on the ground increased, the likelihood of choosing awareness reduced.
222 Specifically, experience most strongly predicted the likelihood of choosing awareness compared to
223 enforcement ($p < 0.01$, Odds Ratio, 0.72, 0.95 CI: 0.58-0.91), or stakeholder ($p < 0.01$, Odds Ratio,
224 0.78, 0.95 CI: 0.66-0.91), or impacts ($p < 0.05$, Odds Ratio, 0.80, 0.95 CI: 0.68-0.95). As the HDI of
225 participants' home nation increased so did the likelihood of choosing stakeholder interventions. This
226 effect was strongest in predicting stakeholder interventions being chosen compared to awareness (p
227 < 0.01 , Odds Ratio, 1.41, 0.95 CI: 1.21-1.61), or enforcement ($p < 0.01$, Odds Ratio, 1.35, 0.95 CI:
228 1.11-1.67), or to a lesser extent, impacts ($p < 0.05$, Odds Ratio, 1.18, 0.95 CI: 1.01-1.37). Participant
229 HDI also predicted the likelihood of choosing economic interventions compared to awareness ($p <$
230 0.01 , Odds Ratio, 1.25, 0.95 CI: 1.06-1.47).

231

232 As respondent age increased the likelihood of choosing both enforcement and awareness reduced.
233 Age most strongly predicted the likelihood of choosing enforcement compared to stakeholder ($p <$
234 0.01 , Odds Ratio, 0.63, 0.95 CI: 0.52-0.76), or economic interventions ($p < 0.01$, Odds Ratio, 0.63,
235 0.95 CI: 0.51-0.77), or to a lesser extent, impacts ($p < 0.01$, Odds Ratio, 0.76, 0.95 CI: 0.63-0.93).
236 Similarly, age predicted the likelihood of choosing awareness compared to economic interventions
237 ($p < 0.01$, Odds Ratio, 0.75, 0.95 CI: 0.67-0.90) or stakeholder interventions ($p < 0.01$, Odds Ratio,
238 0.77, 0.95 CI: 0.68-0.88). Male respondents were more likely than females to prioritise enforcement
239 compared to stakeholder interventions ($p < 0.01$, Odds Ratio, 1.42, 0.95 CI: 1.05, 1.93), but less likely
240 to prioritise awareness ($p < 0.05$, Odds Ratio, 0.64, 0.95 CI: 0.49-0.84), or impacts ($p < 0.05$, Odds

241 Ratio, 0.62, 0.95 CI: 0.43-0.91). In most models, the variation between individual respondents was
242 largest with regards to enforcement (Supporting information Table S2) and generally the models
243 explained a high proportion of the total variation (Model Set 2, mean $R^2 = 0.21$).

244

245 Respondents' also highlighted the importance of local contextual and multi-faceted interventions
246 (often including stakeholder engagement as a starting point). Of the 166 respondents who gave
247 written feedback, 43% described the need, or benefit, of combinations of interventions. In total 30%
248 of respondents asked for more context or described contextual factors which would influence their
249 decisions. However, only 7% mentioned the geographical location or development level of the
250 conflict country, only 2% referenced the legality of behaviours and only 4% commented on the
251 conflicts framing. Moreover, 7% requested information about the species (such as habitat and
252 conservation status). Other interventions which were suggested included hunting (2%), lethal
253 control (2%), and other forms of non-lethal technical interventions (3%). In total, 23% of
254 respondents outlined the need to prioritise stakeholder-based interventions first, to either increase
255 buy-in (6%), better understand a conflict (7%) (including drawing upon community knowledge) and
256 to help tackle the social roots at the heart of conflicts (4%). Only 4% discussed enforcement
257 (Supporting Information, Appendix 2).

258

259 **Discussion**

260 The results of the experimental survey suggest that particular characteristics of wildlife conflicts and
261 the characteristics of decision-makers influence intervention recommendations. Whilst it is known
262 that people with different backgrounds and experiences favour different approaches for
263 conservation generally (Sandbrook et al. 2019) and for conflicts specifically (Lute et al. 2018), this

264 study sheds further light on these differences and highlights the possible processes and factors
265 influencing how conservationists make decisions.

266

267 This study illuminates the importance of contextual cues on conservation decision-making. Relatively
268 simple changes to the objective description of a conflict, such as the conflict location or whether a
269 behaviour is described as illegal or not, had big effects on intervention priorities. Likewise, contexts
270 which appear comparable in terms of the general problem – wildlife impacts and retaliatory killing –
271 and which differed only in terms of taxa, types of competing human interests and types of wildlife
272 impacts, promoted different solutions. Impact reduction efforts for instance are widely prioritised
273 for crop-raiding bears, but are largely overlooked for fish-eating otters or lamb-raiding sea-eagles.
274 Such contextual effects could be generated by numerous mechanisms. For instance, they might
275 represent a form of cognitive bias, reflecting fast, intuitive thinking (Papworth 2017) and the priming
276 effects of specific words (Bargh 2006). Alternatively, they might reflect respondents' values,
277 assumptions and conceptualisations related to their understanding of specific species, countries, or
278 conflict contexts (Game et al. 2013). Indeed, for some respondents, such knowledge and experience
279 (both first-hand and through literature) base might inform more deliberative, reflective decisions
280 (Papworth 2017). Whilst this study doesn't illuminate which processes are dominant here,
281 conservation managers generally more heavily upon experience and intuition than published
282 scientific evidence (Walsh et al. 2015).

283

284 Beyond highlighting the general importance of context, we also identify specific associations
285 between conflict characteristics and intervention decisions. Our finding that enforcement and
286 awareness were favoured more for scenarios situated in more highly developed countries, and by
287 respondents from less highly developed countries was unexpected. In a previous review,
288 enforcement appeared to be more commonly recommended by authors for conflicts in less highly

289 developed nations, and awareness showed no associations (Baynham-Herd, et al. 2018). However,
290 that study was observational and therefore could not account for the additional variation between
291 conflict situations as we did here. Instead, we propose three reasons to account for why
292 enforcement (though generally prioritised the least) was favoured in more developed countries: the
293 possible widespread appreciation of the critiques of militarised and enforcement-based
294 conservation in the Global South (Duffy et al. 2019; Mabele 2017), perceptions that wildlife-related
295 killings are less legitimate in more highly developed countries (Dickman et al. 2015; Sheil et al. 2016)
296 or the understanding that successful enforcement is contingent upon effective governance
297 (Sundström 2015). However, enforcement was infrequently discussed in the written feedback, hence
298 further investigations would be needed to ascertain to what extent different practical and ethical
299 reasons - such as cultural relativism (Dickman et al. 2015) – might account for this effect.

300

301 That economic interventions appeared to be more commonly suggested in less highly developed
302 countries, stands in contrast with the finding that conflict-related compensation is more common in
303 highly developed countries (Ravenelle & Nyhus 2017). However, it is possible that the lack of
304 incentives and compensation schemes in less highly developed nations might be a result of the
305 greater structural challenges in providing them rather than varying priorities (DeMotts & Hoon
306 2012), despite the apparently healthy appetite for them among researchers and practitioners
307 identified here. Survey feedback also hinted at the idea, common in the conservation literature
308 (Salerno et al. 2016), that the material costs of conflicts may be relatively greater in less highly
309 developed nations – such as where food insecurity, or dependence on forest resources is higher.
310 However, the non-material impacts of conflicts are also clearly significant in the Global South (Barua
311 et al. 2013) and the social roots of conservation conflicts are likely to be just as strong between less
312 and highly developed countries (Young et al. 2013). Moreover, given that conservation rule breaking
313 everywhere is frequently associated with acts of resistance and not just material incentives (Holmes

314 2007) we also suggest a need for further investigation into the reasons why economic interventions
315 were prioritised more when wildlife killing was described as illegal. This is particularly important
316 given that conservation payments can also lead to reductions in previously unrewarded positive
317 conservation behaviours (Fisher 2012). The lack of the importance of the conflict framing variable
318 was unexpected, suggesting either different conflict frames are less important than predicted
319 (Baynham-Herd, et al. 2018), or at least less salient than the other factors tested. Further work
320 should explore the extent to which conservation researchers and practitioners might be influenced
321 by perceptions and assumptions made about countries in different stages of development, which are
322 often out-of-date or inaccurate (Rosling & Zhang 2011).

323

324 Our finding that respondent characteristics – such as disciplinary background, age and conflict
325 experience – predict their intervention decisions, highlights the importance of socio-demographic
326 influences on conservation decision making (Papworth 2017). This supports previous findings that
327 conflict management priorities differ across regions and respondents' backgrounds (Lute et al.
328 2018). We suggest that further work should explore whether disciplinary backgrounds and
329 experience of conflicts on the ground shape the way decision-makers conceptualise conflicts – such
330 as the emphasis placed on social relations (Sandbrook et al. 2013).

331

332 Although we cannot provide as clear explanations to account for the apparent effects of age, gender
333 and development status of respondents' home nation, these factors have also been shown to
334 predict conservation priorities more generally. For instance age, gender and regional origin all
335 predict respondent' general conservation rationale and support for market-based conservation
336 (Sandbrook et al. 2019). and gender can predict local management preferences (Keane et al. 2016)
337 and attitudes to particular taxa (Suryawanshi et al. 2014).

338

339 We cannot say from our data whether prioritisations were also influenced by the factors not
340 experimentally manipulated: such as taxa, previous knowledge, actual prevalence or likelihood of
341 each described conflict, impact severity or conservation status. Likewise, although we instructed
342 respondents to ignore the issue of resources, it is possible that perceived differences in management
343 costs (Iacona et al. 2018) may have tacitly influenced decisions. Similarly, although our sample size is
344 appropriate, our conclusions are limited to generalisations about largely Anglo-European sample,
345 which reflects the Anglo-European bias in conservation conflict research (Baynham-Herd, et al. 2018;
346 van Eeden et al. 2018), but doesn't represent other voices in conservation decision-making
347 (Sandbrook et al. 2019).

348

349 These results have important implications for wildlife conflict management. Firstly, if context-
350 contingent intervention priorities, such as those identified here, are informed by reasoned thinking
351 and evidence, they may produce effective outcomes (Sutherland & Wordley 2017). If however, such
352 decisions are more shaped by unknown biases and predispositions, they may not (Papworth 2017).
353 Hence, decision-makers could benefit both from further personal retrospection (identifying their
354 own biases and assumptions) and from further studies which test prevailing assumptions in conflict
355 management (van Eeden et al. 2018). Secondly, given that the characteristics of decision-makers
356 also shape intervention priorities, increasing the diversity of those involved in conflict decision-
357 making would not only be ethical but may improve decisions (Green et al. 2015). For instance,
358 increased female (Leisher et al. 2016), community (Mishra et al. 2017) and interdisciplinary (Bennett
359 et al. 2017) participation, in decision-making has been found to improve a range of conservation
360 outcomes. Furthermore, whilst different conservation managers and stakeholders are unlikely to
361 always agree – for both practical and value-based reasons (Rust 2017; St John et al. 2018) – better
362 understanding other's positions and increasing dialogue helps fostering more effective collaboration

363 (Lute et al. 2018; Game et al. 2013). Thirdly, both the survey results and feedback support recent
364 scholarship (Redpath et al. 2017) in highlighting participatory and stakeholder-first conflict
365 interventions as best-practice and in advocating for multi-pronged (Hazzah et al. 2014) and adaptive
366 management strategies (Bunnefeld et al. 2017). Education and awareness programs were often cited
367 in feedback as being necessary additions to any interventions. However, given the failures of many
368 awareness-based conservation programs (Schultz 2011), a further exploration into why and where
369 conservation decision-makers deem them most appropriate is important. Indeed, more targeted
370 approaches such as social-marketing (Salazar et al. 2018) might be more effective than simple
371 information provision, or indeed – often-problematic – enforcement (Duffy et al. 2019). However,
372 how different interventions compliment, or negate each other, is an area in need of greater
373 exploration by both researchers and practitioners (van Eeden et al. 2018).

374

375

376 **Supporting Information**

377 Further information on the sample (Appendix S1) additional results (Appendix S2), and the full
378 survey (Appendix S3) are available online. The authors are solely responsible for the content and
379 functionality of these materials. Queries (other than absence of the material) should be directed to
380 the corresponding author.

381

382

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<p><input type="checkbox"/> A2</p> <p> Due to a lack of wild prey, the threatened Tibetan wolf of northern Nepal is dependent on livestock for food, leading to conflict between conservationists and farmers, with wolves being occasionally illegally killed in the region.</p> <p>Please select which management option to prioritise</p> <ul style="list-style-type: none"><input type="radio"/> Wildlife impact reduction (e.g. livelihood protection, barriers, relocation)<input type="radio"/> Awareness/training (e.g. livelihood/conservation education or awareness)<input type="radio"/> Enforcement/patrols (e.g. ranger patrols, monitoring, penalties)<input type="radio"/> Incentives/compensation (e.g. compensation, payments, insurance)<input type="radio"/> Stakeholder engagement (e.g., consultations, community relations)	<p><input type="checkbox"/> C2</p> <p> Due to a lack of wild prey, the threatened Iberian wolf of northern Portugal is dependent on livestock for food, leading to human-wildlife conflict, with wolves being occasionally killed in the region.</p> <p>Please select which management option to prioritise</p> <ul style="list-style-type: none"><input type="radio"/> Wildlife impact reduction (e.g. livelihood protection, barriers, relocation)<input type="radio"/> Awareness/training (e.g. livelihood/conservation education or awareness)<input type="radio"/> Enforcement/patrols (e.g. ranger patrols, monitoring, penalties)<input type="radio"/> Incentives/compensation (e.g. compensation, payments, insurance)<input type="radio"/> Stakeholder engagement (e.g., consultations, community relations)
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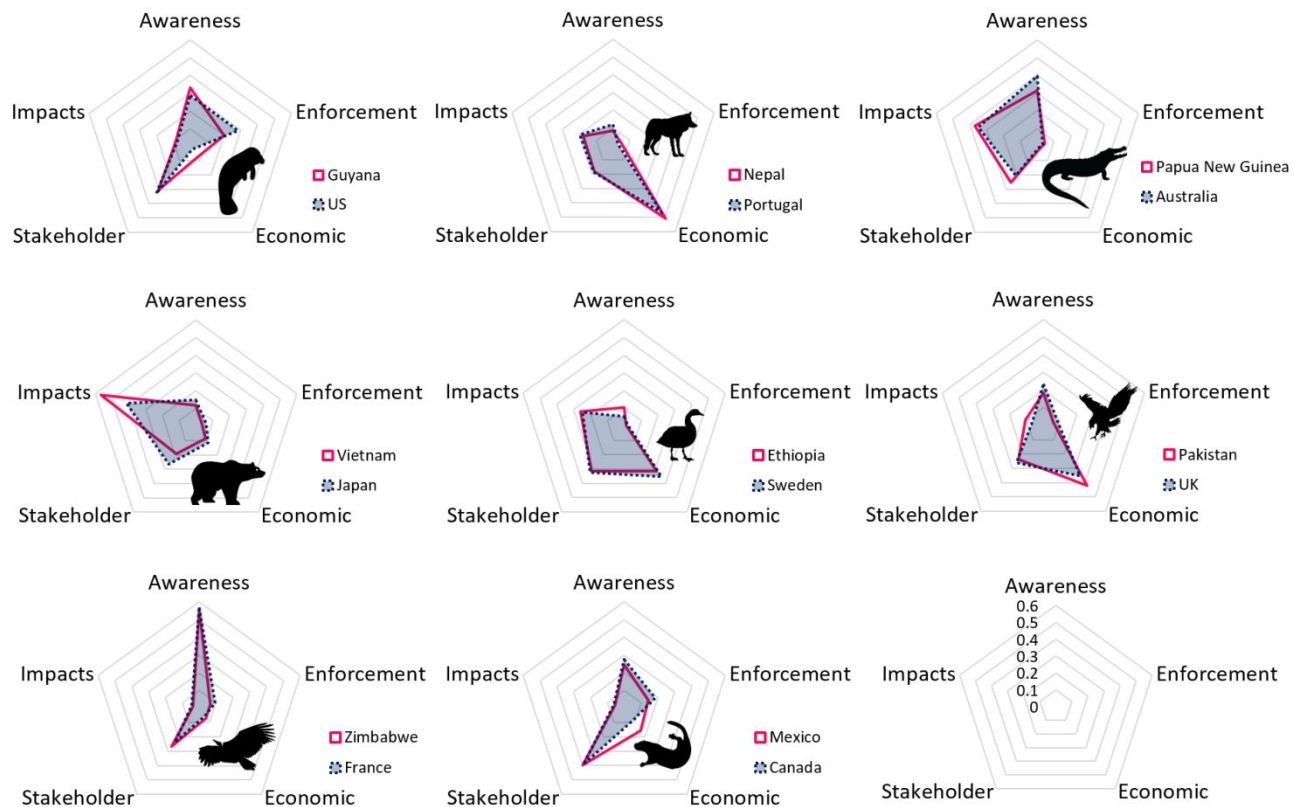
550 **Figure 1:** An example of two different versions of the same scenario which were presented to
551 different participants, from two of eight different question blocks. In this case the location and
552 framing, and the illegality of wildlife killing differs between the two scenarios.

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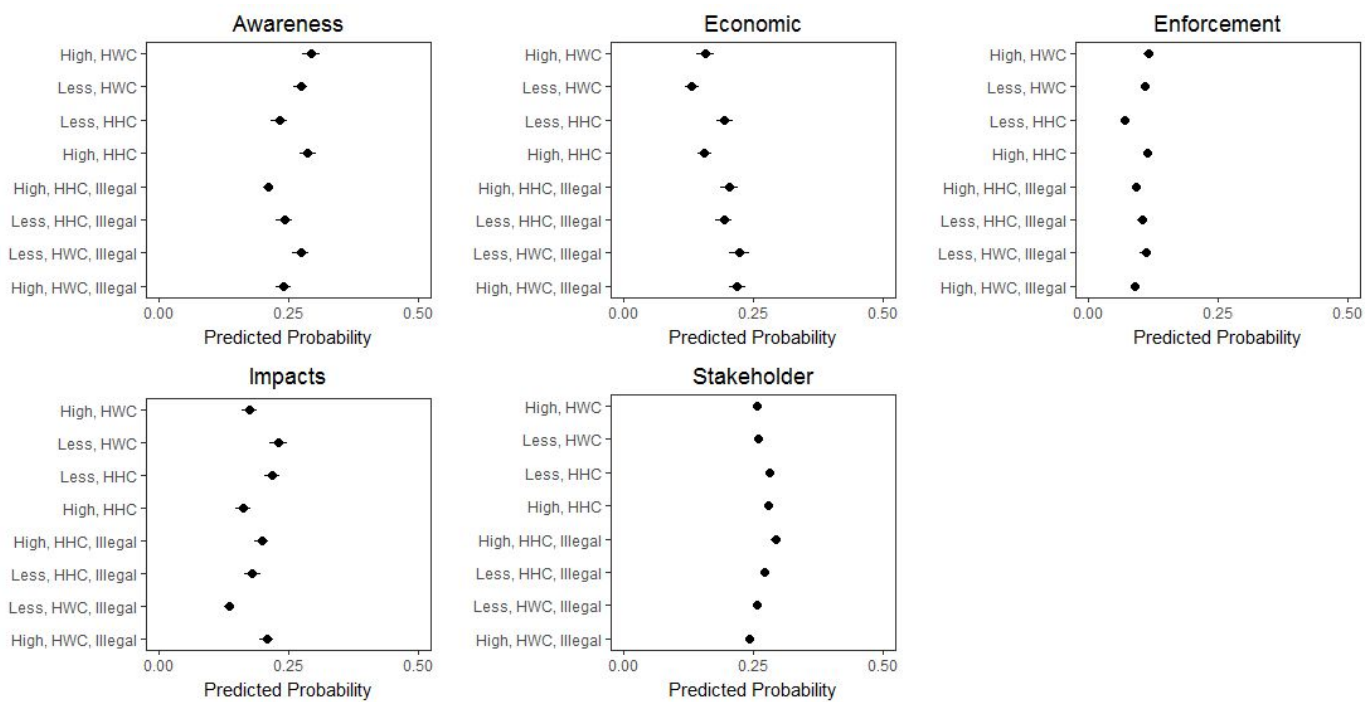


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558 **Figure 2:** Radar charts showing the proportion of different intervention types suggested for each
 559 country in each of the eight conflict scenarios

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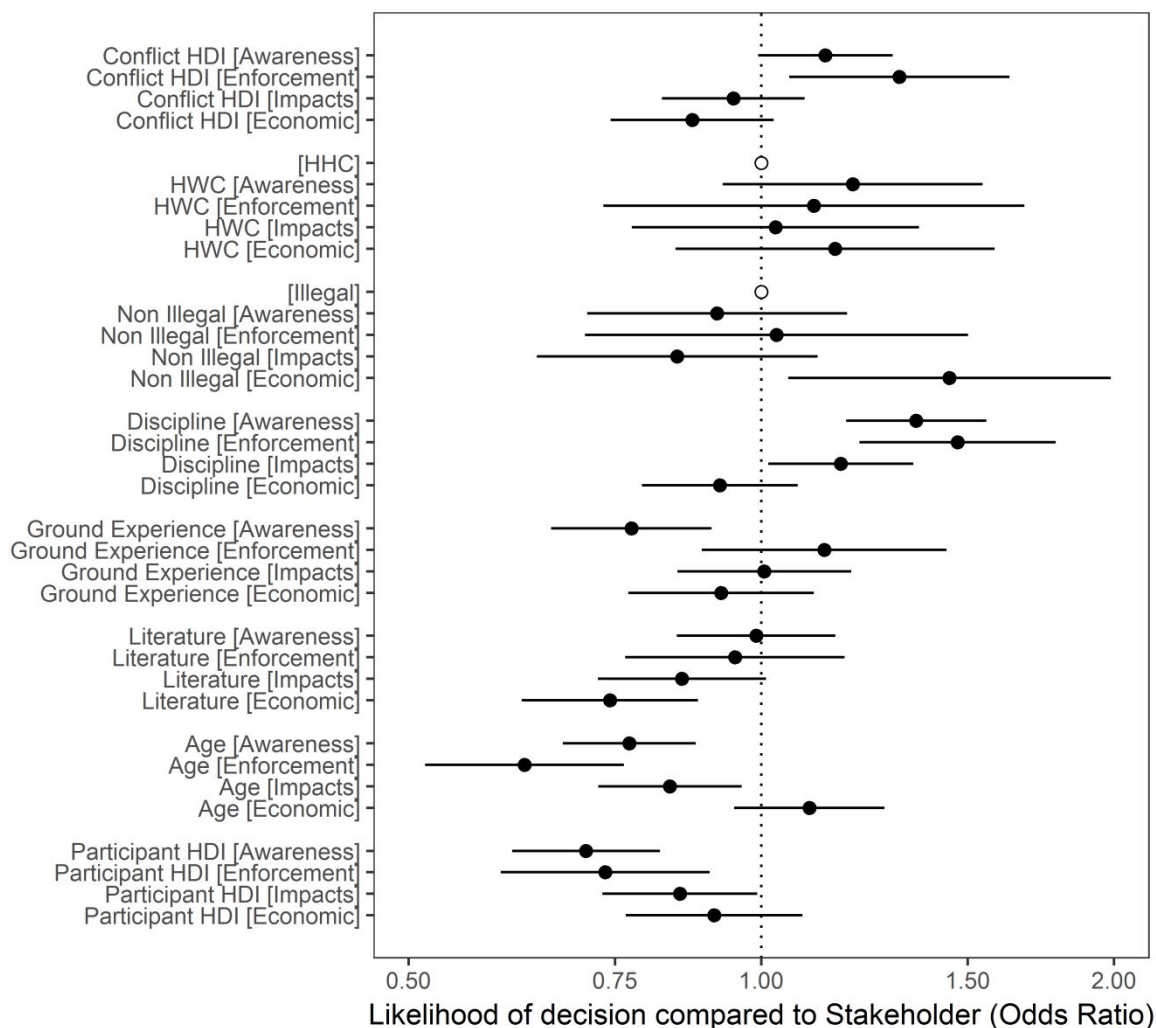
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Figure 3: Results from a multinomial mixed logit regression model (Model Set 1, reference level ‘stakeholder’), showing the predicted probability of choosing each intervention type (panels) under each of the eight framing combinations. Whiskers represent 95% CI. “HWC” = Human-wildlife conflict, “HHC” = Human-human conflict, “High” = “Very highly developed nation”, “Less” = “High, Medium or Less highly developed nation”, “Illegal” = Behaviour described as illegal.



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571 **Figure 4:** Results from a multinomial logit regression model (Model 2), showing the estimated
 572 conditional effects of each predictor variable on likelihood of choosing each intervention type
 573 compared to stakeholder interventions. Filled dots represent model coefficient estimates converted
 574 to odds ratios, which show the expected change in likelihood of a choice when each continuous
 575 variable increases by a unit of one, or when each factor variable changes level from a baseline
 576 (unfilled dots). Whiskers represent 95% CI, and variables with whiskers that do not cross zero are
 577 those predicted by the model to associate with intervention decisions (effect size is distinguishable
 578 from zero). Larger odds ratios indicate greater predicted strength of association. HDI = Human
 579 Development Index, HWC = Human-wildlife conflict frame, HHC = Human-human conflict frame,
 580 Discipline = Disciplinary Background, Literature = Literature knowledge.

581 **Table 1:** A short description of each of the eight conflict scenarios provided in each survey and the
 582 two, systematically rotated, countries they were described as being located in.

Conflict scenario description	Countries	References
American manatee (<i>Trichechus manatus</i>)	USA	(Mason et al. 2018; Solomon et al. 2004; Castelblanco-Martínez et al. 2012)
Conflict between commercial fishing interests and manatee conservation, with manatees drowning in fishing nets and being injured by boats in certain areas with speed restrictions.	Guyana	
Gray wolf (<i>Canis lupus</i>)	Portugal	(Pimenta et al. 2017; Fernández-Gil et al. 2016; Werhahn et al. 2017)
Conflict between rural livestock herding and conservation interests, with wolves predated upon livestock and being killed in retaliation.	Nepal	
Saltwater crocodile (<i>Crocodylus porosus</i>)	Australia	(Fukuda et al. 2015)
Conflict between human safety and conservation interests, with crocodile-related injury and retaliatory killing	Papua New Guinea	

Geese (<i>e.g., Anser anser, Alopochen aegyptiaca</i>)	Sweden	(Tombre et al.
Conflict between agriculture and conservation interests, with crop-raiding and retaliatory scaring or killing	Ethiopia	2013)
Sea eagle (<i>Haliaeetus albicilla</i>)	Scotland	(Marquiss et al.
Conflict between rural livestock farming and conservation interests, with livestock depredation and retaliatory killing	Pakistan	2004)
Vulture (<i>e.g., Gyps fulvus, Gyps africanus</i>)	France	(Margalida et
Conflict between rural livelihoods and conservation interests, with livestock depredation, perceived spread of disease and retaliatory killing	Zimbabwe	al. 2014; Ogada et al. 2016)
Sea otter (<i>Enhydra lutris</i>)	Canada	(Echeverri et al.
Conflict between fishing and conservation interests with competition for catch and associated killing	Mexico	2017; Carswell et al. 2015)
Asiatic black bear (<i>Ursus thibetanus</i>)	Japan	(Can et al.
Conflict between agriculture and human safety and conservation interests with crop-raiding, attacks and retaliatory killing	Vietnam	2014; Takahata et al. 2013)

584 **Table 2:** Descriptive summary of variables used in multinomial mixed logit models, using the sample
 585 of 411 responses in which all eight scenarios (3,288 decisions) were completed.

Explanatory Variables	Levels	Source	Descriptive summary (N)	Model Set
Frame	[Human-Human conflict] (HHC) [Human-Wildlife conflict] (HWC)	Experimental manipulation	Scenarios = HHC (1644), HWC (1644)	1,2
Illegal Behaviour	[Illegal] [Non Illegal]	Experimental manipulation	Scenarios: Illegal (1644), Non illegal (1644)	1,2
Conflict HDI		Experimental manipulation (from UNDP)	Mean = 0.75, SD = 0.17, Range = 0.45-0.94	1,2
Question Block	[A-H]	Survey	Scenarios: A (360), B (520), C (568), D (336), E, (368), F (320), G (408), H (408)	1,2
Scenario	[1-8]	Survey	Scenarios: 411 each	1,2
Disciplinary Background		Survey (subjective scale)	Mean = 75.9, SD = 23.64, Range = 0 (Social Sciences/Humanities only) - 100 (Natural Sciences/Ecology only)	2

Ground		Survey	Mean = 62.16, SD = 26.84,	2
Experience		(subjective scale)	Range = 0 (no experience) - 100 (main specialism)	
Literature		Survey	Mean = 66.07, SD = 22.73,	2
Knowledge		(subjective scale)	Range = 0 (no knowledge) - 100 (main specialism)	
Age		Survey	Mean = 37.92, SD = 10.99, Range = 20-80	2
Participant HDI		Survey	Mean = 0.84, SD = 0.12, Range = 0.42-0.95	2
Gender	[Male] [Female]	Survey	Female (207), Male (197)	3
Career Position	[Early][Mid][Senior]	Survey	Early (180), Mid (112), Senior (109)	-
Role	[Researcher] [Practitioner/Other]	Survey	Researcher (321) Practitioner/Other (84)	-

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591 **Table 3:** Results from multinomial logit regression models (Model Set 2), showing the estimated conditional effects of each predictor variable on the
 592 likelihood of choosing each intervention type compared to the reference level in each model (in brackets), with effects presented as odds ratios (OR)
 593 showing the expected change in likelihood of choosing different interventions when each continuous variable increases by a unit of one, or when each
 594 factor variable changes level from a baseline.^a

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	Awareness [Stakeholder]	Enforcement [Stakeholder]	Impacts [Stakeholder]	Economic [Stakeholder]	Awareness [Enforcement]	Impacts [Enforcement]	Economic [Enforcement]	Awareness [Impacts]	Economic [Impacts]	Awareness [Economic]
HDI		OR 1.31* (1.06-1.63)				OR 0.75** (0.60-0.93)	OR 0.70** (0.56-0.88)	OR 1.18* (1.02-1.36)		OR 1.26** (1.07-1.47)
HWC										
Illegal				OR 1.45* (1.05-1.99)					OR 1.49* (1.07-2.07)	OR 0.66** (0.48-0.89)
Discipline	OR 1.36*** (1.18-1.56)	OR 1.47*** (1.21-1.78)	OR 1.17* (1.01-1.35)				OR 0.75** (0.61-0.91)	OR 1.21* (1.04-1.40)		OR 1.38*** (1.18-1.63)
Ground	OR 0.78** (0.66-0.91)				OR 0.72** (0.58-0.91)			OR 0.80* (0.68-0.95)		
Literature				OR 0.74*** (0.62-0.88)						OR 1.21* (1.01-1.45)
Age	OR 0.77*** (0.68-0.88)	OR 0.63*** (0.52-0.76)	OR 0.84* (0.73-0.96)			OR 1.31** (1.07-1.59)	OR 1.59*** (1.30-1.95)		OR 1.24** (1.06-1.44)	OR 0.75** (0.67-0.90)
Nation	OR 0.71*** (0.62-0.82)	OR 0.74** (0.60-0.90)	OR 0.85* (0.73, 0.99)					OR 0.85* (0.73-0.98)		OR 0.80** (0.68-0.94)

596 ^aThe values in brackets represent 95% CI and larger odds ratios indicate greater predicted strength of association and only significant associated are presented. HDI =

597 Human Development Index, HWC = Human-wildlife conflict frame. *P < 0.05, **P < 0.01, ***P < 0.001

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599 **Figure captions:**

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601 **Figure 1:** An example of two different versions of the same scenario which were presented to different participants, from two of eight different question
602 blocks. In this case the location and framing, and the illegality of wildlife killing differs between the two scenarios.

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608 ""HHC" = Human-human conflict, "High" = "Very highly developed nation", "Less" = "High, Medium or Less highly developed nation", "Illegal" = Behaviour
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611 **Figure 4:** Results from a multinomial logit regression model (Model 2), showing the estimated conditional effects of each predictor variable on likelihood of
612 choosing each intervention type compared to stakeholder interventions. Filled dots represent model coefficient estimates converted to odds ratios, which
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614 from a baseline (unfilled dots). Whiskers represent 95% CI, and variables with whiskers that do not cross zero are those predicted by the model to associate
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617 knowledge.

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For review only