

1 PERSPECTIVE

2 **Special Issue on Invasive Mammal Species**

3 **The challenges of long-term invasive mammal management: lessons**
4 **from the UK**

5

6 Aileen C. MILL *Modelling Evidence and Policy Group, School of Natural and Environmental Sciences,*
7 *Newcastle University, Newcastle, NE1 7RU, UK. Email: aileen.mill@ncl.ac.uk*

8 Sarah L. CROWLEY *Environment and Sustainability Institute, University of Exeter, Penryn Campus,*
9 *Penryn, Cornwall, TR10 9FE, UK. Email: S.L.Crowley@exeter.ac.uk*

10 Xavier LAMBIN *School of Biological Sciences, Tillydrone Avenue, Zoology Building, University of*
11 *Aberdeen, Aberdeen AB24 2TZ, UK. Email: x.lambin@abdn.ac.uk*

12 Conor MCKINNEY *Ulster Wildlife, McClelland House, 10 Heron Road, Belfast, BT3 9LE, UK. Email:*
13 *Conor.McKinney@ulsterwildlife.org*

14 Gwen MAGGS *Saving Scotland's Red Squirrels, Harbourside House, 110 Commercial Street,*
15 *Edinburgh, EH6 6NF, UK. Email: gmaggs@scottishwildlifetrust.org.uk*

16 Pete ROBERTSON *Modelling Evidence and Policy Group, School of Natural and Environmental*
17 *Sciences, Newcastle University, Newcastle, NE1 7RU, UK. Email: peter.robertson@ncl.ac.uk*

18 Nikki J ROBINSON *The Wildlife Trusts, The Kiln, Waterside, Mather Road, Newark, Nottinghamshire,*
19 *NG24 1WT, UK. Email: nrobinson@wildlifetrusts.org*

20 Alastair WARD *Department of Biological and Marine Sciences, University of Hull, Cottingham Road,*
21 *Hull, HU6 7RX, UK. Email: A.I.Ward@hull.ac.uk*

22 Mariella MARZANO *Social and Economic Research Group, Forest Research, Northern Research*
23 *Station, Roslin, Midlothian, EH25 9SY, UK. Email: mariella.marzano@ForestResearch.gov.uk*

24

25 **ABSTRACT**

26 We consider the motivations, strategies and costs involved in invasive mammal management
27 undertaken in the UK. Widespread established invasive mammals require long-term management to
28 limit damage or spread, but ongoing management is costly and complex. Long-term management is
29 most effective where it is applied at a landscape scale, but this requires overarching co-ordination
30 between multiple stakeholders. Five challenges for successful long-term management of invasive
31 mammal species are identified as: defining landscape scale strategies, management co-ordination,
32 stakeholder and community engagement, sustainable funding and evidence requirements. We make
33 recommendations on the supportive infrastructure needed for effective landscape-scale
34 management of invasive mammals to fulfil long-term conservation aims, as follows. 1. There is a
35 need for evidence-based Invasive Species Action Plans to provide strategy for the long-term ongoing
36 management of prioritised species at appropriate scales. 2. Where possible, multispecies
37 approaches to invasive species management should be adopted. 3. Trusted leadership should be
38 identified to take ownership of Action Plans and provide an overarching co-ordination to bring
39 individuals, organisations and funders together. 4. Support for a centralised hub for training, data
40 and knowledge flows will greatly improve scientific outcomes through a searchable evidence base
41 and best practise and knowledge sharing.

42 **Key words:** impact, invasive species, landscape, mink, squirrel, stakeholder, United Kingdom

43 **Running Head:** Long-term landscape-scale invasive management

44 Received: 25 April 2019

45 Accepted: 6 November 2019

46 Editor: DR

47 Special Issue Guest Editor: Sandro Bertolino

48

49 INTRODUCTION

50 Invasive species are identifiable by their environmental, economic and societal impacts. Throughout
51 the globe, invasive mammals cause substantial biodiversity impacts, including via predation of native
52 species (Holmes et al. 2019) and disruption to native ecosystems. In the UK, some invasive mammals
53 are so widespread that they are more abundant than native species (Mathews et al. 2018) and have
54 significant ecosystem impacts. The most notable and abundant invasive mammal species in the UK
55 are four deer species (Muntjac *Muntiacus reevesi*, Chinese water deer *Hydropotes inermis*, sika deer
56 *Cervus nippon* and fallow deer *Dama dama*), the American mink *Neovison vison*, and the grey
57 squirrel *Sciurus carolinensis*. All six species have caused the widespread alteration of ecosystems
58 through grazing pressure or serious declines in native species. Mink are responsible for the collapse
59 of populations of water voles *Arvicola amphibius*, terns *Sterna hirundo* and gulls *Laridae* (Craik
60 1997), and grey squirrels have caused the disappearance of the red squirrel *Sciurus vulgaris* in much
61 of England and Wales. Damage from deer and squirrels can nullify economic returns from forestry,
62 with profound long-term consequences for England's rural areas. Without some management, the
63 damage and impacts would remain unmitigated.

64 International targets and policy commitments to manage invasive species such as the Convention for
65 Biological Diversity include commitments to "introduce measures that prevent the introduction and
66 significantly reduce the impacts of [invasive alien species], and control or eradicate priority species"
67 (Convention for Biological Diversity 2014). National strategies for managing the threat of invasive
68 species focus on strengthening biosecurity approaches to prevent new species from entering and,
69 where they do appear, producing contingency plans to ensure their removal before populations can
70 become established. In the UK, there are contingency plans for the removal of newly established
71 invasive mammal species based on the European Union regulation; recent examples include the
72 removal of a racoon *Procyon lotor* in 2016 and two coatimundi *Nasau nasua* in 2018. While
73 eradication of all invasive mammals would remove their impacts, most are so widespread over huge
74 landscapes, that eradication feasibility is limited by cost and practicalities (Booy et al. 2017). When
75 prioritising invasive species for control or eradication, widespread mammal populations are often
76 given low priority in the UK.

77 Managing widespread invasive species on the UK mainland involves ongoing control of local
78 populations by landowning interests to limit local economic damage (such as deer), or control
79 undertaken by stakeholders in the same way as the pest management of native or naturalised
80 species. However, some invasive mammals also cause significant damage to wider biodiversity
81 assets, warranting the involvement of multiple stakeholders (including conservation groups and
82 broader public groups) in the ongoing management of these species.

83 Invasive mammal control projects that do not have eradication as their ultimate objective face
84 multiple challenges. The objectives of ongoing control are often ill-defined, but they generally aim to
85 maintain areas free of species' impacts. Achieving this invariably has significant ongoing costs and

86 requires co-ordination of management efforts over sufficiently large areas to minimise the rate of
87 reinvasion from the periphery. The landscapes invaded by established species are large and can pose
88 challenges for managing, often highly dispersive, invasive mammals. Such landscapes involve
89 multiple stakeholders who have differing priorities around the need for and type of management.
90 This means different groups may be working separately on the same, or different but similarly
91 problematic, species with varying management objectives.

92 The benefit of ongoing control can be limited if it is performed on a small scale and requires
93 continued effort and resources. The co-ordination of such efforts across larger landscapes can
94 achieve enormous benefits. Recognition of this by the New Zealand government has led to an
95 ambitious programme (Predator Free NZ 2050), which aims to remove eight species of non-native
96 mammals from the entire country by 2050 (Russell et al. 2015). Such a level of ambition, and
97 matching financial and community support for ongoing invasive species management as part of the
98 biosecurity agenda, is currently lacking in the UK (Environmental Audit Committee 2019).

99 **MANAGING WIDESPREAD INVASIVE MAMMALS IN THE UK**

100 The principles required to achieve species eradication have been well documented (Bomford &
101 O'Brien 1995) and a framework to assess eradication feasibility has been established (Booy et al.
102 2017). The majority of documented invasive mammal eradications are from oceanic islands and 57%
103 involved removing rodents through poisoning (Jones et al 2016), which is not feasible on larger
104 landmasses where poison may threaten native or domestic species. Larger mammals are most often
105 controlled through trapping or shooting, but the logistics and manpower required are significant and
106 the process can be protracted. There have been documented attempts to eradicate five invasive
107 mammals from mainland UK since the 1930s (Table 1). These programmes have been instigated and
108 co-ordinated by government agencies, and management has typically been undertaken by employed
109 staff rather than volunteers. Eradications of muskrat *Ondatra zibethicus*, coypu *Myocastor coypus*
110 and Himalayan porcupine *Hystrix brachyuran* were implemented to prevent potential economic
111 losses from agricultural damage. These were among the largest successful mammal eradications
112 worldwide, but required a substantial effort over many years to achieve (Robertson et al. 2017).

113 There are currently 12 invasive mammal species being managed across mainland UK (Table 2), and
114 all are too widespread for eradication to be a feasible management objective. Instead, long-term
115 management objectives encompass limiting spread, reducing densities to limit damage or
116 maintaining areas clear of the species. Within defined areas, different control strategies are
117 implemented in response to the management objectives (Fig. 1). Where densities are high, intensive
118 control can be used to reduce impacts to within tolerable thresholds. Where densities and impacts
119 of species are within a tolerable range, routine control is needed to limit density and minimise
120 impacts. At densities lower than the impact threshold, monitoring and surveillance with targeted
121 action can prevent impacts or allow spatial expansion of the protected area.

122 Specifying what density thresholds are tolerable is a socio-ecological question involving
123 consideration of the resilience of native species to invasive species, the economic costs of damage,
124 and the costs of the management required to maintain the density below the threshold. Such costs
125 typically rise non-linearly as density decreases (Hone et al. 2017). Furthermore, the ability of native
126 species and ecosystem functions to be maintained in the presence of invasive mammals is highly
127 variable. Different species and economic activities have different density-impact functions on native
128 biota (Norbury et al. 2015).

129 Three high-profile projects in the UK have recently demonstrated the potential and ambition for
130 greater levels of co-ordinated control, and have convinced funders that this action has sufficient

131 societal benefits to warrant financing over 3-4 years (see Box). The species being managed,
132 American mink and grey squirrel, are widespread with significant biodiversity impacts. These funded
133 projects have given consideration and proposed plausible solutions to the challenges known to
134 confront such open-ended management schemes. However, in contrast with past eradication
135 programmes, these projects are co-ordinated by wildlife charities, not by national government staff,
136 and action is undertaken by a mixture of paid staff, volunteers and community groups (Table 2).
137 Despite having long-term aims, funding is time-limited and projects cannot continue without
138 alternative sustainable resourcing.

139 Here we argue that landscape-scale, co-ordinated approaches to the management of established
140 invasive mammals are necessary, to improve the success rates and cost-effectiveness of current
141 approaches. In reviewing the current landscape-scale invasive mammal management programmes in
142 the UK, we outline five challenges that need to be addressed to improve long-term invasive species
143 management.

145 **Table 1.** Known co-ordinated management approaches of invasive mammal populations in the UK. NGO = non-government organisation.

Management	Species	Population managed	Co-ordination of control	References
National eradication				
	Muskrat <i>Ondatra zibethicus</i>	Successful eradication of 4388 individuals by trapping in Scotland, Ireland, Surrey, Shropshire and Sussex in the 1930s.	Government staff	Warwick (1934, 1940), Sheail (1988), Gosling & Baker (1989)
	Coypu <i>Myocastor coypus</i>	Successful removal of 34822 animals from 19210 km ² in East Anglia, 1980s	Government staff	Gosling & Baker (1989)
	American mink <i>Neovison vison</i>	Failed attempt at national eradication by trapping in 1960s	Government staff	Thompson (1971), Sheail (2004)
	Edible dormouse <i>Glis glis</i>	Failed attempt to eradicate original population	Landowners	Vesey-Fitzgerald (1936)
	Himalayan porcupine <i>Hystrix brachyura</i>	Successful eradication of 6 individuals from SW England in 1970s	Government staff	Baker (2010)
Island eradication				
	American mink <i>Neovison vison</i>	Ongoing eradication attempt across 3461 km ² in Outer Hebrides, Scotland since 2000s	Government staff	Lambin (2014), Roy et al. (2015)
	Grey squirrel <i>Sciurus carolinensis</i>	Successful eradication removed 6397 from 710 km ² on Anglesey, North Wales in 2000s	NGO staff	Schuchert et al. (2014), Shuttleworth et al. (2015)
	Black rat <i>Rattus rattus</i>	Successful eradications from islands up to 4.5 km ² : Lundy, SW England 2004, Shiant, NW Scotland 2018	NGO staff	Appleton et al. (2006), RSPB (2018)
	Brown rat <i>Rattus norvegicus</i>	Successful eradication from islands up to 4.5 km ² : Lundy, SW England 2004, Ailsa Craig, W Scotland 1992, Cardigan SW Wales 1968, Puffin Island, Wales 1998, Ramsey SW Wales 2000, Handa, NW Scotland 1997, Canna W Scotland 2008, Saint Agnes, Gugh, Annet, Isles of Scilly 2016.	NGO staff	Appleton et al. (2006), Zonfrillo (2001), Dryfed Wildlife Trust (1994), Johnstone et al. (2005), Bell et al. (2011), DIISE (2015)
	Feral goat <i>Capra hircus</i>	Failed attempt to eradicate from Holy Island, W. Scotland in 1963	NGO staff	DIISE (2015)

Rabbit <i>Oryctolagus cuniculus</i>	Successful removal from Eynhallow, Orkney, N Scotland 1955, Scolt Head, E England 1955 and Fidra, E Scotland 1961	NGO staff	DIISE (2015)
Regional population removal			
American mink <i>Neovison vison</i>	Ongoing maintenance of cleared area (29000 km ²) in Scottish Highlands	NGO staff and volunteers	Bryce et al. (2011), Melero et al. (2018), Lambin at al. (2018)
Grey squirrel <i>Sciurus carolinensis</i>	Unsuccessful campaign to keep < 46km ² Thetford Forest, East Anglia, clear, 2209 animals removed.	Volunteers	Gurnell & Steele (2002)
Bounty schemes (1950s)			
Coypu <i>Myocastor coypus</i>	Government support for local control groups prior to eradication	Landowners	Sheail (2003)
Grey squirrel <i>Sciurus carolinensis</i>	Nationwide campaign removed one million animals 1953-58	Landowners	Sheail (2004)
Rabbit <i>Oryctolagus cuniculus</i>	Government support for local clearance groups	Landowners	Sheail (1991)

146 **Table 2.** Current on-going management of widespread invasive mammals where objectives are to limit spread, reduce densities to limit damage or maintain
147 areas clear of the species. Estimated UK populations from Mathews et al. (2018) shown with 95% confidence intervals. NGO = non-government
148 organisation. * Brown Hare while introduced is also regarded as a naturalised and is a Biodiversity Action Plan priority species, however is it still controlled
149 for damage to agricultural crops.

Species	estimated UK population	Co-ordination of control
Reeves muntjac <i>Muntiacus reevesi</i>	128,000 (115,000 -147,000)	Landowners
American mink <i>Neovison vison</i>	~122,000	Landowners, NGOs and local volunteer communities
Grey squirrel <i>Sciurus carolinensis</i>	2,700,000 (1,340,000 - 3,790,000)	Landowners, NGOs and local volunteer communities
Edible dormouse <i>Glis glis</i>	23000 (9,800 - 82,000)	Landowners and home-owners
Brown rat <i>Rattus norvegicus</i>	~7,070,000	Landowners and home-owners
Sika deer <i>Cervus nippon</i>	103,000 (27,000 - 266,000)	Landowners
Fallow deer <i>Dama dama</i>	264,000 (194,000 - 343,000)	Landowners

Chinese water deer <i>Hydropotes inermis</i>	3600 (200 - 43,000)	Landowners
House mouse <i>Mus musculus</i>	~5,203,000	Landowners and home-owners
Brown hare* <i>Lepus europaeus</i>	579,000 (427,000 - 1,990,000)	Landowners
Feral goat <i>Capra hircus</i>	Discrete Populations	Landowners
Rabbit <i>Oryctolagus cuniculus</i>	~3,600,000	Landowners

150

151

153 CHALLENGE 1: DEFINING LANDSCAPE-SCALE MANAGEMENT STRATEGIES

154 Removal of widespread and highly dispersive invasive mammals from small areas can be inefficient
155 or futile, as recolonisation from the periphery is inevitable and requires constant monitoring and
156 management (Lawton & Rochford 2007). Funding for repeated invasive species control across the
157 same geographic areas year on year is difficult to obtain, yet often some action is necessary to
158 mitigate damaging impacts. Management strategies that encompass a broader geographic scale are
159 more cost-effective (Robertson et al. 2017), can take account of natural boundaries of suitable
160 habitats (Goldstein et al. 2016) and geographic barriers (such as elevation or coastlines), and can
161 help reduce the likelihood or frequency of reinvasion of a cleared area (Robertson et al. 2018).
162 Defining the appropriate ecologically meaningful scale depends on the species and its interaction
163 with the landscape. Habitat quality can be used as an indicator of suitability, and therefore likelihood
164 of colonisation or key dispersal routes. Targeted control across a landscape focuses on intercepting
165 dispersing mammals or using areas of highly suitable habitat as 'ecological traps' (Melero et al.
166 2018). By integrating knowledge of the ecology of the target species and managing at a sufficient
167 scale, a spatial strategy can be employed to increase the efficiency of control efforts and maximise
168 protection of biodiversity interests across whole landscapes.

169 CHALLENGE 2: CO-ORDINATION OF MANAGEMENT

170 Since 2006, UK wide invasive species policy is set out in the Non-Native Species Framework,
171 delivered by a Secretariat. The Secretariat, however, only consists of a few staff, and their primary
172 role is to provide advice and support for activities such as risk assessment; they are not responsible
173 for decision-making and there is no provision for managing species on the ground. The GB Non-
174 Native Species Framework promotes long-term management and local action through Invasive
175 Species Action Plans and regional Local Action Groups. Much of the current Local Action Group
176 network is focussed on weeds and riparian systems, but work is geared towards co-ordinated action
177 and sharing of best practice. There are currently no Invasive Species Action Plans for mammals;
178 developing these could be the first step towards building co-ordinated approaches. However,
179 realisation of these plans would require co-ordination between different projects, and agreement as
180 to how each project contributes to the aims of the action plan.

181 The GB Non-Native Species Framework recognises that the development and implementation of a
182 GB strategy should build on existing capacity and expertise to achieve effective co-ordination of
183 existing management activities. Some statutory bodies and organisations are responsible for invasive
184 species control on land they own or manage. However, they have no responsibility for co-ordinating
185 action at a national scale. Some species-specific initiatives and networks exist with varying objectives
186 and structures. The Deer Initiative, a broad partnership of organisations with interests in deer
187 management, all of whom are signatory to an accord, and whose operations are delivered by a small
188 team of employees working to achieve consensus among landholders (www.thedeerinitiative.co.uk),
189 has in part been a victim of its own success. Improvements to woodland condition (a common
190 objective across landscapes) are evident in areas where the Deer Initiative has been facilitating
191 collaboration and co-ordination of stakeholder action at landscape scales for several years. However,
192 changes to funding streams and the transferability of its activities to its partner organisations mean
193 that the Deer Initiative will close from March 2020. Whether metrics of deer management
194 performance continue to indicate success in the absence of the Deer Initiative remains to be seen.
195 The UK Squirrel Accord is a partnership between a number of agencies and organisations with an
196 interest in grey squirrel management (<http://squirrelaccord.uk/>); the Northern Ireland Squirrel
197 Forum is a similar partnership of statutory and non-statutory organisations in Northern Ireland.

198 These partnerships do not provide funding for control, and currently do little to co-ordinate and
199 guide management for a common objective. A management co-ordination role, e.g. Saving
200 Scotland's Red Squirrels (see Box), should include data sharing and directing management efforts to
201 high priority areas as well as facilitating shared best practise, equipment and centralised training.

202

203 CHALLENGE 3: STAKEHOLDER CONTRIBUTIONS AND COMMUNITY ENGAGEMENT

204 There is increasing recognition of the importance of stakeholder involvement in research and
205 management of invasive species and that increased interaction between policy-makers, academics
206 and practitioners can improve management approaches (Pages et al. 2019; Shackleton et al. 2019).
207 Landowner support is crucial to a landscape-scale approach to invasive mammal management, but
208 in some places access to land can be a barrier to ensuring comprehensive coverage. Lack of
209 landowner commitment may be due to other priorities, limited resources or because of conflicting
210 views or the fear of public opposition to controversial management of charismatic species (Niemiec
211 et al. 2019). Where landowners are supportive of the management objectives of invasive species
212 control, access requests to trap or shoot can be refused on the grounds of health and safety
213 concerns or insurance requirements. Demonstration of training and adherence to standards can
214 alleviate these concerns (e.g. professionally accredited courses), and many local volunteer groups
215 fundraise to cover personal and liability insurance and training needs of their members. Having a
216 landscape-scale strategy can help to achieve buy-in from neighbouring landowners. Awareness
217 raising, relationship development and advocacy are important for knowledge exchange and
218 dissemination, and for gaining trust and support between stakeholders of invasive mammal control
219 campaigns. However, short funding timeframes do not provide sufficient time to allow these
220 relationships to be built. Co-ordinating partnerships (see Box) ensures standardised training and
221 accreditation, brings financial benefits of scale, and are a good way to ensure appropriate methods
222 and high welfare standards are maintained.

223 In the UK, there are precedents for small volunteer-based groups leading conservation efforts. For
224 example, there are >30 registered local red squirrel conservation groups working throughout the NE
225 of England, many facing similar resourcing challenges. Researchers have identified multiple drivers
226 for environmental volunteering, including pro-social motivations such as contributing to their
227 community, personal reasons such as increasing social interactions or developing skills, a general
228 ethic of care towards the environment and/or a particular attachment to the place or issue
229 concerned (Measham & Barnett 2008). Volunteers are involved in invasive species management in
230 several ways, including monitoring, fundraising, public engagement and direct assistance with
231 control. Volunteer effort can also vary from the relatively simple and infrequent to more challenging
232 and time-consuming roles, such as co-ordinating a volunteer group or contributing specialist skills to
233 a management scheme. Although there is often a diversity of roles available to volunteers, the most
234 popular activities may be oversubscribed (e.g. monitoring red squirrels in areas where they are
235 found), while more mundane or challenging tasks can be difficult to recruit to (e.g. few people are
236 willing to kill animals or to gain the appropriate training).

237 Despite the benefits of recruiting engaged, skilled and motivated volunteers, heavy reliance on
238 volunteer effort can nevertheless pose challenges for invasive species management projects. First,
239 these initiatives often rely on the ongoing engagement of people who may be time-poor, and whose
240 other commitments may understandably come before their voluntary work. Second, volunteer
241 retention can be challenging unless volunteers see rapid results or feel their continued efforts are
242 making a difference. Unfortunately, the timescales over which landscape-scale management

243 strategies make observable change are long, and even where progress is being made this may not be
244 immediately apparent to volunteers. Volunteering as an extension of existing work may be one way
245 to ensure long-term participation; for example, fisheries employees have additionally volunteered
246 on mink trapping programmes over extended periods, as the volunteering duties fit reasonably well
247 with the objectives and schedules of their daily work (Lambin et al.2018).

248 To fully engage stakeholder participation in landscape-scale control and to determine what
249 management action is feasible, there is a need to combine the motivations and knowledge of the
250 conservation and stakeholder community with academic understanding of both social and ecological
251 environmental processes. This integration can be achieved with support and expertise from
252 academic and non-governmental organisations (Reed 2008), but it requires co-ordination and
253 sustained funding.

254 **CHALLENGE 4: SUSTAINABLE FUNDING**

255 Obtaining funding for repeated invasive species control in the same geographic areas year on year is
256 difficult, yet action is necessary to mitigate damaging impacts. Where there is no co-ordination of
257 long-term management, limited funding can unwittingly be spent resourcing the same infrastructure
258 and project support needs, e.g. designing data collection platforms, and implementing training and
259 public awareness-raising campaigns, often to the detriment of action on the ground. The lack of
260 sustainable funding can also result in competition between organisations working on the same
261 species, who may lose sight of the bigger picture and rarely work towards a common strategy.

262 The majority of funding for invasive mammal management projects is currently through EU LIFE,
263 (<https://ec.europa.eu/easme/en/life>) and the National Lottery Heritage Fund
264 (<https://www.heritagefund.org.uk/>), which typically finance projects over three to five years. These
265 funders are increasingly looking for novel projects that engage with volunteers and local
266 communities. This is a laudable approach, but there is concern that money is being detracted from
267 direct action, leading to tensions over what takes priority when it comes to invasive mammal
268 management. Funders rarely support follow-on projects, forcing the continual rebranding and
269 formulation of project aims for what is essentially on-going control.

270 Funded co-ordinated action for long-term control at the UK governmental level is lacking, as
271 widespread species are of low priority and funding is not committed to open-ended projects.
272 Without central funding, maintaining ongoing control through phased projects is both challenging to
273 achieve and damaging to project identity and retention of a workforce (Lambin et al. 2019).
274 Alternative funding sources are needed to support the co-ordination of activities under a
275 recognisable banner with a long-term commitment, such as the strategy of Saving Scotland's Red
276 Squirrels. Access to alternative funding sources or incentives may result from Brexit or from
277 Common Agricultural Policy reforms, but to achieve landscape-scale objectives, we need a better
278 mechanism to support projects across species and between areas to achieve the best impact
279 (Environmental Audit Committee 2019).

280 **CHALLENGE 5: EVIDENCING PROGRESS AND BENEFITS**

281 Data collation is essential to demonstrate the success of invasive mammal management, but also
282 offers the opportunity for learning and sharing of knowledge gained. A key feature of scaling action
283 across landscapes is the need to collate and share data at these larger scales, to record action and
284 document progress (Lambin et al. 2018). These data may simply be mapping activity or may include
285 more detailed records of surveillance effort, control action and biodiversity or economic gains that
286 can be used to assess the impact of control efforts formally, and to plan future action (Harrington et
287 al. 2020).

288 Evaluation of management effort is critical to securing public funding for grey squirrel control (Bryce
289 & Tonkin2018). The Saving Scotland’s Red Squirrels project (see Box) has launched a web-based
290 community hub (<https://scottishsquirrels.org.uk/>) aimed at collecting evidence to support the
291 project whilst providing an online resource for data recording and information. The aim of the hub is
292 to connect interested members of the public, volunteers and landowners across a large country, and
293 create a centralised location for the project’s data and volunteer management. The hub
294 accommodates activities for the public through reporting and viewing squirrel sightings, for
295 registered volunteers through monitoring and trapping data, and for independent network groups to
296 co-ordinate all their activities and data entry. The hub also allows volunteers to track their training,
297 and record volunteer hours, and promotes Scotland-wide networking. While digital innovation is the
298 way forward, there is a danger that, without co-ordination, different projects invest and design their
299 own individual data-collecting mechanisms, making data sharing and comparison at a broad scale
300 more challenging or impossible (Wilson et al. 2018). A recent Scottish Biodiversity Information
301 Forum review of data recording recognised a need for a professional function that focuses on
302 mobilising and collecting data, not just for invasive species, but across all aspects of conservation
303 (Wilson et al. 2018). A co-ordinated approach may make it possible to harness data from
304 consultancies, developers and private businesses who may be carrying out invasive species control in
305 some areas yet do not feed into existing mechanisms.

306 Strategic management programmes use modelling approaches to improve efficiencies and the
307 spatial deployment of effort (Melero et al. 2018). Landscape-scale models of habitat suitability and
308 population dynamics can help prioritise management action. However, these approaches require
309 understanding of species dispersal and habitat requirements to be able to predict where in the
310 landscape species present the greatest risk, or knowledge of population parameters, such as survival
311 and fecundity rates, to assess the level and type of management required to prevent further
312 population growth (Ward et al. 2020).

313 Often, the optimal strategy for controlling or monitoring invasive species is not known at the outset
314 and a degree of flexibility within the strategy is needed to allow learning about the system to be
315 integrated into the management. Adaptive management approaches can improve the effectiveness
316 of management, e.g. through comparison of observed impacts of different spatial or temporal
317 trapping strategies with predictions from population models that include management (Bryce et al.
318 2011). The framework requires an agreed measurable objective (e.g. quantified reduction in
319 population density) and systematic monitoring of the management effectiveness to update
320 understanding of the system and consequently refine the management action. To date, monitoring
321 and adaptive management processes have been under-utilised in invasive species management
322 (Richardson et al. 2020), but a common strategic approach could facilitate their uptake and improve
323 the cost-efficiency of programmes (Will et al. 2014). Adaptive management can also be used to
324 integrate multiple management approaches, such as the use of the latest technology and
325 techniques. For example, automated remote traps (Jones et al. 2015) can reduce effort, as they do
326 not need a manual check every day, and the use of conservation dogs can increase species
327 detection.

328 CONCLUSIONS

329 Funding for invasive mammal species control in the UK has been decreasing steadily over the years,
330 leading to competition between conservation-based organisations over funding to manage species
331 in their own areas. Mammal control in the UK relies on the dedication of conservationists and
332 landowners to reduce or eliminate the impacts of invasive species, but only large-scale efforts offer
333 the chance of achieving sustainability. Working at ecologically relevant scales maximises the use of

334 landscape geography to manage potential reinvasion, and offers economies of scale to make best
335 use of expertise and volunteering efforts. While there are inherent challenges in working at larger
336 scales, co-ordination across landscapes, organisations, volunteers and species is feasible. If
337 achieved, the potential benefits would be enormous and the potential for species eradications may
338 become reality. Currently, we are faced with limited or short-term funding that inhibits project
339 effectiveness, longevity and relationship building. Individual projects are often localised, species-
340 specific, and reliant on the good will, motivation and skills of volunteers that also entail management
341 and training transaction costs.

342 Joined-up approaches could facilitate a standardised way to collect data that will improve
343 significantly the ability to compare scientific data, prioritise management actions, evidence impact
344 and offer a mechanism to share knowledge and lessons learned. Co-ordination will also provide a
345 centralised source of training in invasive species management to maintain high welfare standards
346 and offer economies of scale, so that resources can be spent on actual management such as control
347 activities. However, joining up organisations, stakeholders and volunteers across species and
348 landscapes will require careful co-ordination to provide all the benefits of the large scale, whilst
349 allowing for local contexts and characteristics. If we are to improve the chances of species and
350 ecosystems under threat in the UK, it is time for relevant bodies to acknowledge the inefficiencies
351 and challenges of current approaches, despite considerable efforts at local scales, and identify a
352 sustainable pathway going forward for species management. With this in mind, we recommend the
353 following principles as a benchmark for future discussions on joined-up thinking and action.

- 354 1. There is a need for evidence-based Invasive Species Action Plans to provide strategy for the
355 long-term ongoing management of prioritised species at appropriate scales.
- 356 2. Where possible, multispecies approaches to invasive species management should be
357 adopted.
- 358 3. Trusted leadership should be identified to take ownership of Action Plans and provide an
359 overarching co-ordination to bring individuals, organisations and funders together.
- 360 4. Support for a centralised hub for training, data and knowledge flows will greatly improve
361 scientific outcomes through a searchable evidence base and best practise and knowledge
362 sharing.

363 ACKNOWLEDGEMENTS

364 We thank all participants at the Red Squirrels United panel discussion on landscape-scale
365 management at the Mammal Society's Autumn Symposium on Invasive Species in November 2018.

366 REFERENCES

- 367 Appleton D, Booker H, Bullock DJ, Cordrey L, Sampson B (2006) The seabird recovery project: Lundy
368 Island. *Atlantic Seabirds* 8: 51-59.
- 369 Baker SJ (2010) Control and eradication of invasive mammals in Great Britain. *Revue Scientifique Et*
370 *Technique-Office International Des Epizooties* 29: 311-327. doi:10.20506/rst.29.2.1981.
- 371 Bell E, Boyle D, Floyd K, Garner-Richards P, Swann B, Luxmoore R, Patterson A, Thomas R (2011)
372 *Island Invasives: Eradication and Management*. International Union for Conservation of Nature and
373 Natural Resources (IUCN), Gland, Switzerland.
- 374 Bomford M, O'Brien P (1995) Eradication or control for vertebrate pests. *Wildlife Society Bulletin* 23:
375 249-255.
- 376 Booy O, Mill AC, Roy HE, Hiley A, Moore N, Robertson P *et al.* (2017) Risk management to prioritise
377 the eradication of new and emerging invasive non-native species. *Biological Invasions* 19: 2401-
378 2417. doi:10.1007/s10530-017-1451-z.

379 Bryce J, Tonkin JM (2019) Containment of invasive grey squirrels in Scotland: meeting the challenge.
380 In: Veitch C, Clout M, Martin A, Russell J, West C (eds) *Proceedings of the Island Invasives: Scaling up*
381 *to Meet the Challenge*, 180–186. IUCN Gland, Switzerland.

382 Bryce R, Oliver MK, Davies L, Gray H, Urquhart J, Lambin X (2011) Turning back the tide of American
383 mink invasion at an unprecedented scale through community participation and adaptive
384 management. *Biological Conservation* 144: 575-583. doi:10.1016/j.biocon.2010.10.013.

385 Convention for Biological Diversity (2014) *Pathways of Introduction of Invasive Species, their*
386 *prioritization and management*. Convention on Biological Diversity.

387 Craik C (1997) Long-term effects of North American mink *Mustela vison* on seabirds in western
388 Scotland. *Bird Study* 44: 303-309. doi:10.1080/00063659709461065.

389 DIISE. (2015) The Database of Island Invasive Species Eradications. IUCN SSC Invasive Species
390 Specialist Group,, Coastal Conservation Action Laboratory UCSC, University of Auckland and
391 Landcare Research New Zealand.

392 Dyfed Wildlife Trust (1994) *Fifty Years- Cardigan Island*. Dyfed Wildlife Trust Bulletin No. 65.

393 Environmental Audit Committee (2019) Invasive species. First Report of Session 2019 Published by
394 House of Commons 2019.

395 Goldstein EA, Butler F, Lawton C (2016) Modeling future range expansion and management
396 strategies for an invasive squirrel species. *Biological Invasions* 18: 1431-1450. doi:10.1007/s10530-
397 016-1092-7.

398 Gosling L, Baker S (1989) The eradication of muskrats and coypus from Britain. *Biological Journal of*
399 *the Linnean Society* 38: 39-51. doi:10.1111/j.1095-8312.1989.tb01561.x.

400 Gurnell J, Steele J (2002) *Grey squirrel control for red squirrel conservation: a study in Thetford forest*.
401 English Nature Research Report 453, English Nature, Peterborough, UK

402 Harrington L, Birks J, Chanin P, Tansley D (2020) Current status of American mink *Neovison vison* in
403 Britain: a review of the evidence for a national-scale population declines. *Mammal Review*: In Press.

404 Holmes ND, Spatz DR, Opper S, Tershy B, Croll DA, Keitt B *et al.* (2019) Globally important islands
405 where eradicating invasive mammals will benefit highly threatened vertebrates. *PLoS ONE* 14: 17.
406 doi:10.1371/journal.pone.0212128.

407 Hone J, Drake VA, Krebs CJ (2017) The effort-outcomes relationship in applied ecology: evaluation
408 and implications. *BioScience* 67: 845-852. doi:10.1093/biosci/bix091.

409 Johnstone IG, Gray CM, Noble DG (2005) *The state of birds in Wales 2004*. RSPB Cymru, Cardiff, UK.

410 Jones C, Warburton B, Carver J, Carver D (2015) Potential applications of wireless sensor networks
411 for wildlife trapping and monitoring programs. *Wildlife Society Bulletin* 39: 341-348.
412 doi:10.1002/wsb.543.

413 Jones HP, Holmes ND, Butchart SH, Tershy BR, Kappes PJ, Corkery I *et al.* (2016) Invasive mammal
414 eradication on islands results in substantial conservation gains. *Proceedings of the National Academy*
415 *of Sciences* 113: 4033-4038. doi:10.1073/pnas.1521179113.

416 Lambin X, Cornulier T, Oliver M, EJ. F (2014) *Analysis and future application of Hebridean Mink*
417 *Project data*. Scottish Natural Heritage, Commissioned Report No. 522

418 Lambin X, Horrill JC, Raynor R (2019) Achieving large scale, long-term invasive American mink control
419 in northern Scotland despite short term funding. In: Veitch CR, M.N. C, Martin AR, J.C. R, West CJ
420 (eds) *Proceedings of the Island Invasives: Scaling up to Meet the Challenge*. IUCN, Gland, Switzerland.

421 Lawton C, Rochford J (2007) The recovery of grey squirrel (*Sciurus carolinensis*) populations after
422 intensive control programmes. *Biology and Environment-Proceedings of the Royal Irish Academy*
423 107B: 19-29. doi:10.3318/bioe.2007.107.1.19.

424 Mathews F, Kubasiewicz LM, Gurnell J, Harrower CA, McDonald RA, Shore RF (2018) *A Review of the*
425 *Population and Conservation Status of British Mammals. A Report by the Mammal Society Under*
426 *Contract to Natural England, Natural Resources Wales and Scottish Natural Heritage*. Natural
427 England, Peterborough, UK.

428 Measham T, Barnett G (2008) Environmental volunteering: motivations, modes and outcomes.
429 *Australian Geographer* 39: 537-552. doi:10.1080/00049180802419237.

430 Melero Y, Cornulier T, Oliver MK, Lambin X (2018) Ecological traps for large-scale invasive species
431 control: predicting settling rules by recolonising American mink post-culling. *Journal of Applied*
432 *Ecology* 55: 1769-1779. doi:10.1111/1365-2664.13115.

433 Niemiec RM, Willer R, Ardoin NM, Brewer FK (2019) Motivating landowners to recruit neighbors for
434 private land conservation. *Conserv Biol* 33: 930-941. doi:10.1111/cobi.13294.

435 Norbury GL, Pech RP, Byrom AE, Innes J (2015) Density-impact functions for terrestrial vertebrate
436 pests and indigenous biota: guidelines for conservation managers. *Biological Conservation* 191: 409-
437 420. doi:10.1016/j.biocon.2015.07.031.

438 Pages M, Fischer A, van der Wal R, Lambin X (2019) Empowered communities or "cheap labour"?
439 Engaging volunteers in the rationalised management of invasive alien species in Great Britain.
440 *Journal Environmental Management* 229: 102-111. doi:10.1016/j.jenvman.2018.06.053.

441 Reed MS (2008) Stakeholder participation for environmental management: a literature review.
442 *Biological Conservation* 141: 2417-2431. doi:10.1016/j.biocon.2008.07.014.

443 Richardson S, Mill A, Davis D, Jam D, Ward A (2020) A systematic review of adaptive wildlife
444 management for the control of invasive, non-native mammals and other human-wildlife conflicts.
445 *Mammal Review*: in press.

446 Robertson PA, Adriaens T, Lambin X, Mill A, Roy S, Shuttleworth CM, Sutton-Croft M (2017) The
447 large-scale removal of mammalian invasive alien species in Northern Europe. *Pest Management*
448 *Science* 73: 273-279. doi:10.1002/ps.4224.

449 Robertson PA, Roy S, Mill AC, Shirley M, Adriaens T, Ward AI, Tatayah V, Booy O (2019) Invasive
450 species removals and scale – contrasting island and mainland experience. In: Veitch C, Clout M,
451 Martin A, Russell J, West CJ (eds) *Proceedings of the Island Invasives: Scaling up to Meet the*
452 *Challenge*, 651–657 IUCN, Gland, Switzerland.

453 Roy SS, Chauvenet ALM, Robertson PA (2015) Removal of American mink (*Neovison vison*) from the
454 Uists, Outer Hebrides, Scotland. *Biological Invasions* 17: 2811-2820. doi:10.1007/s10530-015-0927-y.

455 RSPB (2018) *Shiant Isles Recovery Project* [https://ww2.rspb.org.uk/our-](https://ww2.rspb.org.uk/our-work/conservation/shiantisles/work/index.aspx)
456 [work/conservation/shiantisles/work/index.aspx](https://ww2.rspb.org.uk/our-work/conservation/shiantisles/work/index.aspx)

457 Russell JC, Innes JG, Brown PH, Byrom AE (2015) Predator-Free New Zealand: Conservation Country.
458 *BioScience* 65: 520-525. doi:10.1093/biosci/biv012.

459 Schuchert P, Shuttleworth CM, McInnes CJ, Everest DJ, Rushton SP (2014) Landscape scale impacts
460 of culling upon a European grey squirrel population: can trapping reduce population size and
461 decrease the threat of squirrelpox virus infection for the native red squirrel? *Biological Invasions* 16:
462 2381-2391. doi:10.1007/s10530-014-0671-8.

463 Shackleton RT, Adriaens T, Brundu G, Dehnen-Schmutz K, Estevez RA, Fried J *et al.* (2019)
464 Stakeholder engagement in the study and management of invasive alien species. *Journal of*
465 *Environmental Management* 229: 88-101. doi:10.1016/j.jenvman.2018.04.044.

466 Sheail J (1988) The extermination of the muskrat in inter-war Britain. *Archives of Natural History* 15:
467 155-170.

468 Sheail J (1991) The management of an animal population - changing attitudes towards the wild
469 rabbit in Britain. *Journal of Environmental Management* 33: 189-203. doi:10.1016/s0301-
470 4797(05)80095-7.

471 Sheail J (2003) Government and the management of an alien pest species: a British perspective.
472 *Landscape Research* 28: 101-111.

473 Sheail J (2004) The mink menace: the politics of vertebrate pest control. *Rural History* 15: 207-222.

474 Shuttleworth C, Schuchert P, Everest D, McInnes C, Rushton S, Jackson N (2015) Developing
475 integrated and applied red squirrel conservation programmes: what lessons can Europe learn from a
476 regional grey squirrel eradication programme in North Wales? In: Shuttleworth C, Lurz P, Hayward M
477 (eds) *Red Squirrels: Ecology, Conservation and Management in Europe*, 233-250. European Squirrel
478 Initiative, Woodbridge, UK.

479 Thompson HV (1971) British wild mink - challenge to naturalists. *Agriculture* 78: 421-&.

480 Vesey-Fitzgerald B (1936) Welcome or unwelcome guest. *The Field*: 1075.

481 Ward A, Richardson S, Macarthur R, Mill A (2020) Using and communicating uncertainty for the
482 effective control of invasive non-native species. *Mammal Review*: In Press.
483 Warwick T (1934) The distribution of the muskrat (*Fiber zibethicus*) in the British Isles. *Journal of*
484 *Animal Ecology* 3: 250-267.
485 Warwick T (1940) A contribution to the ecology of the musk-rat in the British Isles. *Proceedings of*
486 *the Zoological Society of London A* 110: 165-201.
487 Will DJ, Campbell KJ, Holmes ND (2014) Using digital data collection tools to improve overall cost-
488 efficiency and provide timely analysis for decision making during invasive species eradication
489 campaigns. *Wildlife Research* 41: 499-509. doi:10.1071/wr13178.
490 Wilson E, Edwards L, Judge J, Johnston C, Stroud R, McLeod C, Bamforth L (2018) *A Review of the*
491 *Biological Recording Infrastructure in Scotland by the Scottish Biodiversity Information Forum:*
492 *Enabling Scotland to be a global leader for biodiversity.* Scottish Biodiversity Information Forum
493 Commissioned Report No. 1.
494 Zonfrillo B (2001) *Ailsa Craig before and after the eradication of rats in 1991.* Ayrshire Bird Report,
495 2000.

496

Box – management of American mink and grey squirrels in the UK

Red Squirrels United led a panel discussion at the Mammal Society Symposium Autumn Symposium on Invasive species in November 2018 on the future of invasive mammal management in the UK, with case studies of two invasive mammals - grey squirrel and mink – and their current funded management programmes in the UK. The absence of secured continuous funding for these projects for what are undoubtedly long-term conservation endeavours remains a major challenge to sustaining existing gains or expanding further.

Red Squirrels United. For many years, local red squirrel conservation groups had formed throughout the UK with little overarching co-ordination. Red Squirrels United is a four-year project funded through EU LIFE 14 (LIFE14/NAT/UK/000467) and the UK's National Lottery Heritage Fund from 2016-2020. Led by The Wildlife Trusts, the partnership consists of Northumberland Wildlife Trust, Red Squirrels Trust Wales, Ulster Wildlife, The Wildlife Trust for Lancashire, Manchester and North Merseyside, The Wildlife Trust of South and West Wales, Forest Research and Newcastle University. The Red Squirrels United project works in England, Wales and Northern Ireland and is the first project to bring together grey squirrel management for red squirrel conservation in different countries. The project aims are to support and encourage effective grey squirrel control effort, and to disseminate and evolve best-practice techniques, while protecting and expanding red squirrel populations.



Scottish Invasive Species Initiative. Long-term conservation efforts focussed on Scottish rivers started in 2004, later becoming the Scottish Mink Initiative (2010-2014) and then Scottish Invasive Species Initiative (2017-2021). The mink's northward spread has been stopped, and its ecological impact has been much reduced since

2010 in over 23000 km², 33% of the land mass of Scotland. The defining feature of the project is working with volunteers and local river trusts who have an economic stake in the health of river ecosystems. The initiative uses a formal adaptive management approach focussing on both ecological and socio-ecological issues (reviewed by Lambin et al. 2019). Since 2017, the initiative co-ordinates the management of multiple invasive species, having added five invasive plants to its remit alongside mink. Even though it was part funded by Scottish Natural Heritage, the project expanded, under multiple 3-4 year-long reincarnations with ever more ambitious objectives, over 15 years in the absence of long-term funding.

Saving Scotland's Red Squirrels. The Scottish Wildlife Trust lead a partnership (with Scottish Natural Heritage, Forestry Commission Scotland, RSPB Scotland, Scottish Land & Estates and the Red Squirrel Survival Trust) committed to red squirrel conservation through Saving Scotland's Red Squirrels over a 30-year timescale, with the intention to review the need to continue the work thereafter. The project is currently funded for five years by the National Lottery Heritage Fund (£2.46M



2017-2022). Saving Scotland's Red Squirrels is currently the largest project, geographically, tackling landscape-scale grey squirrel control for the conservation of red squirrels. The success of the project in stabilising and increasing red squirrel populations across strategically selected parts of Scotland has

led to the current phase of developing community action. This new phase aims to engage and empower local communities to take ownership of their local red squirrels, taking on vital project roles for red squirrel protection, creating a project legacy and assisting long-term sustainability.

