

1 **OPINION**

2 **Can biomass supply meet the demands of BECCS?**

3 Michael B. Jones¹ and Fabrizio Albanito²

4 ¹ Botany Department, School of Natural Sciences, Trinity College Dublin, University of
5 Dublin, Dublin 2, Ireland

6 ² Institute of Biological and Environmental Sciences, University of Aberdeen, 23 St Machar
7 Drive, Aberdeen, AB24 3UU, UK

8 Correspondence: Michael B. Jones, Botany Department, School of Natural Sciences, Trinity
9 College Dublin, Dublin 2, Ireland. Email: mike.jones@tcd.ie

10

11

12 **Abstract**

13 In order to reach the reduced carbon emission targets proposed by the Paris agreement one
14 of the widely proposed decarbonizing strategies, referred to as negative emissions
15 technologies (NETs), is the production and combustion of bioenergy crops in conjunction
16 with carbon capture and storage (BECCS). However, concerns have been increasingly raised
17 that relying on the potential of BECCS to achieve negative emissions could result in delayed
18 reductions in gross CO₂ emissions, with consequent high-risk of overshooting global
19 temperature targets. We focus on two particular issues; the carbon efficiency and payback
20 time of bioenergy use in BECCS and the potential constraints on the supply of bioenergy.
21 The simplistic vision of BECCS is that one ton of CO₂ captured in the growth of biomass
22 equates to one ton of CO₂ sequestered geologically, but this cannot be the case as CO₂ is
23 emitted by variable amounts during the life cycle from crop establishment to sequestration
24 below ground in geological formations. The deployment of BECCS is ultimately reliant on the
25 availability of sufficient, sustainably sourced, biomass. The two most important factors
26 determining this supply are land availability and land productivity. The upper bounds of the
27 area estimates required correspond to more than the world's harvested land for cereal
28 production. To achieve these estimates of biomass availability requires the rapid evolution
29 of a multitude of technological, social, political, and economic factors. Here, we question
30 whether, because of the limited sustainable supply of biomass, BECCS should continue to be
31 considered the dominant NET in IPCC and other scenarios achieving the Paris targets, or
32 should it be deemed no longer fit for purpose?

33

34 **1. Introduction**

35 Negative emissions technologies (NETs) are any process that removes carbon dioxide from
36 the atmosphere and stores it in the biosphere or geosphere. In recent years, the
37 international research on NETs has grown rapidly and publications have ranged in scope
38 from reviewing the potential of NETs in climate change mitigation scenarios, to assessing
39 the feasibility of achieving technological maturity and discussing deployment opportunities
40 (Minx et al., 2017). However, concerns have been increasingly raised that ungrounded
41 optimism in NETs potential could result in delayed reductions in gross CO₂ emission, with
42 consequent high-risk of overshooting of global temperature targets (Kato & Yamagata,
43 2014; Fuss et al., 2014; Anderson & Peters, 2016; Vaughan et al., 2018). In order to reach
44 the reduced carbon emission targets proposed by the Paris agreement, one of the widely
45 proposed NETs is the production and combustion of forest products or second-generation
46 bioenergy crops in conjunction with carbon capture and storage (BECCS). Most Integrated
47 Assessment Models (IAMs) suggest that BECCS will make a significant contribution to NETs
48 in the near to mid-term future and it has come to be viewed as the key CO₂ removal
49 approach to keep global atmospheric CO₂ concentrations below 500 ppm and avoid
50 catastrophic climate change (Fuss et al., 2018; Nemet et al., 2018; NAS, 2018). The reason
51 why BECCS plays such a pervasive and pivotal role in climate change mitigation pathways is
52 based on an assumption that large areas of land could be made available for bioenergy
53 production at scale and that bioenergy is in the near term a relatively low-cost and low-
54 emission source of energy (EASAC, 2018; Reid, Ali & Field, 2019).

55 Negative emissions as a consequence of BECCS is achieved when the CO₂ absorbed from the
56 atmosphere during the growth cycle of biomass is released in combustion and energy
57 production and then captured and stored indefinitely (Kemper, 2015). The simplistic vision
58 of BECCS is that one ton of CO₂ captured in the growth of biomass equates to one ton of CO₂
59 sequestered geologically. However, biomass crops are not carbon neutral because GHG
60 emissions are associated with the cultivation of biomass and furthermore GHG emissions
61 occur throughout the BECCS value chain which reduces the carbon efficiency (Brandao et al.,
62 2018; Tanzer & Ramirez, 2019).

63 Gough et al. (2018) identified a number of policy and governance challenges associated with
64 the deployment of BECCS, including whether BECCS can be delivered at sufficient scale and
65 also be provided sustainably. Here we suggest that it is becoming increasingly clear that the
66 potential of BECCS is significantly constrained by a combination of socio-political, technical
67 and geographic considerations, including limits to knowledge and experience (Fridahl &
68 Lehtveer, 2018), and ask whether BECCS should continue to be the dominant NET in IPCC
69 and other scenarios achieving Paris targets or should it be deemed no longer fit for purpose.
70 We highlight two particularly important issues, the low carbon efficiency and long payback
71 time of biomass for BECCS and the potential constraints on the supply of biomass for
72 bioenergy.

73

74

75

76

2. Carbon efficiency of BECCS

77

78

79

80

81

82

83

84

85

86

87

88

89

90

91

92

The first impact of displacing fossil fuel with biomass for power generation in BECCS is a carbon debt derived from an increase in atmospheric CO₂ emission from land clearing and biomass production (*i.e.* direct land use emissions), harvesting, transport and processing (*i.e.* lifecycle emissions), as well as stack emissions due to higher CO₂ emissions from biomass combustions relative to continued fossil fuel use. The size of the bioenergy carbon debt depends also on how far upstream and downstream emissions are estimated within the system boundary of any life cycle analysis (LCA) (Tanzer & Ramirez, 2019). Studies including only a gate-to-gate boundary system ignore land-based emissions and assume that the amount CO₂ removed by biomass from the atmosphere is equal to the CO₂ emitted from its combustion (*i.e.* the bioenergy is “carbon neutral”). These studies in general provide optimistic results on the potential savings from bioenergy crops. Since bioenergy systems involve land-based and lifecycle emissions, a further expansion of the boundaries is needed to encompass a “cradle-to-gate” boundary system which includes upstream emissions (Hastings, 2017, Roder et al., 2015). Ultimately, LCAs should include both upstream and downstream emissions in a “cradle-to-grave” determination of the overall climate mitigation of NETs (Tanzer & Ramirez 2019) (Figure 1).

93

94

95

96

97

98

99

100

101

102

103

104

105

106

107

108

109

Life cycle GHG assessments have shown that the differences in supply chain emissions from biomass cultivation, harvesting and transportation can be relatively small compared to the large differences in combustion and processing efficiencies of the power plants (Odeh & Cockerill 2008, Roder et al., 2015). This is because, at the point of combustion, biofuels generate more CO₂ per unit of end-use energy than fossil fuels. Although woody biomass has approximately the same carbon intensity as coal (0.027 vs. 0.025 tC/GJ of primary energy), combustion efficiency of wood and wood pellets is in general lower than fossil fuels with typical combustion efficiencies for wood being approximately 25%, compared to 35% for coal (IEA 2016). Published estimates vary with the process examined, but energy processing losses for the wood pellet supply chain are on the order of approximately 27% if biomass is used in the drying process (Roder et al., 2015), compared to losses of approximately 11% for coal (IEA, 2016). In addition, capturing the CO₂ and then compressing it prior to transport produce an energy penalty that needs to be accounted for in the overall boundary system of BECCS. Additional energy is required to extract the steam needed for the CO₂ absorber/stripper system, to scrub the CO₂ due to compressing the flue gas and pumping the solvent, and finally to compress the recovered CO₂ prior to belowground storage sequestration.

110

3. Carbon payback time

111

112

The carbon payback time of biofuels is the number of years it can take to offset the carbon emissions generated by converting land for biofuels. Biofuels can only reduce

113 atmospheric CO₂ over time by increasing net primary productivity above what it otherwise
114 would have been. This period is very sensitive to the type of biomass crop and previous land
115 use, and can range from less than 10 years for perennial grasses to over 100 years for slow
116 growing clear-felled forests. Estimates of the payback time of biomass-fired power plants
117 are limited, but have been reported as high as 400 years (Bentsen 2017). Depending on the
118 land use change, the land-based production system and fossil fuel replaced, the payback
119 time of the carbon debt of wood pellets is found to vary from 44 to >104 years to offset coal
120 (Sterman et al., 2018). Pellets from residues are reported to have a temporal lags against
121 coal-fired generation of approximately 16 years, while pellets from standing tree harvesting
122 would require 35-50 years to reach a payback (McKechnie et al., 2011). It is also important
123 to recognise that BECCS negative emissions are not delivered from year one and a time lag
124 will occur before the initial extra emissions from producing the crop and establishing the
125 BECCS facility are recovered.

126

127 **4. Impacts of land use for BECCS**

128 Land carbon stocks are influenced both by direct land use change (dLUC) involved in
129 switching to the BECCS crop and from secondary impacts (*e.g.* in shifting demand for food to
130 new areas) which lead to indirect land use change (iLUC). It is particularly difficult to
131 summarize the potential emissions from iLUC, as its impact and extent is a function of
132 spatial ecology, macro-scale economics, and type of biomass feedstock considered. So that
133 iLUC can only be modelled and not measured directly. Second generation biofuels have
134 been reported to have a median carbon loss of 5 gCO₂-eq/MJ, although sequestration is
135 possible mainly in marginal areas, where perennial crops (such as switchgrass (*Panicum*
136 *virgatum*), miscanthus (*Miscanthus x giganteus*) and *Arundo donax*) can increase below-
137 ground carbon content leading to negative emissions of -12 gCO₂-eq/MJ (Valin et al., 2015).
138 Pellets from forest residues have been reported to have a mean loss of 17 gCO₂-eq/MJ
139 biofuel due to lower build-up of soil organic carbon stocks (Valin et al., 2015). Straw and
140 stover could have losses varying from 2 to 3 gCO₂-eq/MJ biofuel (Overmars et al., 2015),
141 while for cereal straw they vary from 0 to 16 gCO₂-eq/MJ biofuel (Valin et al., 2015) (see
142 Table 1).

143 **5. Balancing biomass supply and demand**

144 The deployment of BECCS is ultimately reliant on the availability of sufficient, sustainably
145 sourced, biomass for an active CCS industry operating at scale and a favourable policy and
146 commercial environment to incentivise these investments (Boysen, Lucht & Gerten, 2017).
147 IAM scenarios aimed at keeping warming below 2°C include global demand for sustainable
148 biomass for BECCS ranging from 60 EJ/year up to more than 500 EJ/year (Fuss et al., 2018).
149 The supply of biomass needs to be sufficient to either provide for centralised power stations
150 or distributed energy systems, such as combined heat and power (CHP) stations. There are

151 also opportunities to capture CO₂ from other sources such as ethanol biorefineries using
152 first generation crops (Sanchez et al., 2018). The two most important factors determining
153 the biomass supply for BECCS are land availability and land productivity. These factors are in
154 turn determined by competing uses of land and a myriad of environmental and economic
155 considerations (Searle & Malins, 2014). Published estimates of the potential for biomass
156 supply vary widely, due mainly to the mixture of methodologies, assumptions and datasets
157 employed (Batidzirai et al., 2012). Different estimates derive in part from differing utilisation
158 of the same data collated by the Food and Agriculture Organisation (FAO), and from the use
159 of different assumption on the spatial and temporal factors affecting biomass potential
160 production. Furthermore, because in the FAO data-set there is a paucity of data for some
161 regions, such as Africa, the lack of robust, reliable and high resolution information make
162 global spatial biomass assessments difficult to model.

163 The potential sources of biomass for BECCS range from harvested residues from first
164 generation food crops to forests, managed short-rotation coppice such as willow and
165 poplar, and second generation biomass crops such as perennial rhizomatous grasses. The
166 advantage of the second generation crops is that they can produce usable energy with less
167 than 10% the energy inputs of first generation food crops, and lower water and nutrient
168 requirements. However, the viability of achieving the highest yields demonstrated in
169 experimental plots over large areas and many different types of soils has not yet been
170 demonstrated. For example Searle & Mallins (2015) reviewed the yields of five major energy
171 crops (*Miscanthus*, Switchgrass, Poplar, Willow, and Eucalyptus) all of which had produced
172 high yields in small, intensively managed trials. However, yields were significantly lower in
173 semi-commercial scale trials, due largely to biomass losses with drying and harvesting
174 inefficiency under real world conditions.

175 Recent studies provide a large range for the global technical potential of biomass supply for
176 2050, ranging from ~30 to over 1,000 EJ/yr (Table 1). However the higher levels are
177 considered implausible either because the estimates of available land are too optimistic or
178 yield expectations are inflated by extrapolations from pilot-based studies to large areas of
179 less productive marginal and severely degraded land (Smith & Torn, 2013; Smith et al.,
180 2014). Furthermore the ability of degraded and marginal land to produce economic yields
181 without the use of expensive fertilisers and irrigation is vastly overestimated (Field,
182 Campbell & Lobell, 2007). In addition, relating the unevenly distributed biomass supply to
183 the amount of CO₂ that can be stored geologically (Hendricks et al., 2004) is complex as the
184 available regional information does not account for the possibility that CO₂ storage capacity
185 may also be unevenly distributed within a given region (Muri, 2018).

186 The global land areas needed for the deployment of BECCS by 2100 have been estimated to
187 range from 380 to 700 Mha (Smith et al., 2015). The upper bounds correspond to three
188 times the world's harvested land for cereal production, twice the current water use for
189 agriculture and 20 times the current US annual fertiliser use. The reason why the range is so

190 large, revolves around differences in the definition of marginal land that is abandoned or
191 severely degraded. Also, while there is evidence that the area of abandoned land has been
192 increasing, with the FAO Landsat analysis indicating that across the tropics 77 Mha of
193 cropland and pasture had been abandoned either temporarily or permanently during the
194 1990s (FAO, 2001; Gibbs et al., 2010), the productivity of this land is very low.

195 As well as abandoned land, it seems that globally there is a sizable area of land that is used
196 for agriculture but which makes a small contribution to food production and could be used
197 for biomass production (approximately 600 Mha) (Alexandratos and Bruinsma, 2012, Mason
198 et al., 2015). Considering only low productivity cropland with terrain slope $\geq 20\%$ (i.e.
199 disadvantaged agricultural areas), Albanito et al. (2016) suggested that the global land
200 suitable for bioenergy would be approximately 250 Mha, which could generate up to 318 EJ
201 of primary energy.

202 Spatially explicit studies on the availability of biomass for BECCS are limited to the USA and
203 UK. Baik et al., (2018) reported that by 2020 in the US up to 230 Mt of lignocellulosic
204 biomass could be available annually producing as much as 400 Mt CO₂/year for
205 sequestration in BECCS. The UK has a theoretical storage CO₂ capacity of 78 GtCO₂ with 50%
206 confidence (Bentham et al., 2014), and previous research reported that BECCS could
207 mitigate between 4.5 and 55 MtCO₂/year and approximately 1.5 Mha could be made
208 available to bioenergy crops (ETI, 2016; Smith et al., 2016). However, considering only the
209 low quality grade agricultural land available to produce an environmentally and
210 economically sustainable supply of biomass to the power sector, Albanito et al. (2019)
211 reported that only 0.4 to 0.5 Mha would be available across Great Britain.

212 **6. Are biomass for BECCS ambitions sustainable?**

213 In recent years the widely accepted criteria for a sustainable biomass supply have evolved to
214 ensure that the biomass: (i) is not redirected from food or animal feed purposes, (ii) does
215 not reduce the ecological functioning of the land, (iii) is grown on marginal land not suitable
216 or economically attractive for food crop production, and (iv) is utilised locally (within ~50
217 km) to limit transport costs (Lewandowski, 2015).

218 Smith et al., (2015) make it clear that bioenergy systems deployment needs to balance a
219 range of environmental, social, and economic objectives that are not always fully
220 compatible. The effectiveness and consequences of bioenergy development depends on: (i)
221 the technology used, (ii) the location, scales, and pace of implementation (iii) the land used,
222 *i.e.* forest, grassland, crop lands, and marginal land, including how they displace existing
223 land use, and (iv) the business models and practices adopted. The conclusion is that
224 estimates of availability for the future depends on the evolution of a multitude of social,
225 political, and economic factors including land tenure and regulation, trade, and technology
226 (Smith et al., 2015; Bui et al., 2018).

227 An additional and rapidly emerging issue is the realisation that there will be competing
228 demands for biomass to provide the feedstock of a rapidly growing bio-economy (Dahmen
229 et al., 2019). Using biomass for energy is a high volume, low value operation while biomass
230 used for the production of bio-based chemicals and materials, which will be essential if the
231 demands to remove non-degradable plastics from the environment are to be met, has a
232 much higher value.

233 We suggest that further work is thus required to quantify the sustainable capabilities for
234 BECCS as a NET. This should demonstrate that, before being given priority in future climate
235 change reduction strategies, the risks can be managed effectively through not only technical
236 means but also international governance and the impacts that NETs will have on sustainable
237 development goals and equity issues between nations (Fajady & Mac Dowell, 2017;
238 Fuhrman et al., 2019). Moreover, energy policy should not overlook the inherently low
239 efficiency of exploiting photosynthesis (the basic process driving conversion of CO₂ to
240 biomass) for energy, since the amount of electricity which can be produced from a hectare
241 of land using Photo Voltaic (PV) is at least 50-100 times that of biomass (Baldocchi &
242 Penuelas, 2018). Reducing uncertainty in the outcomes is crucial to increase the robustness
243 of decisions that use integrated assessment models as inputs and more sensitivity analyses
244 should be made in order to understand the implications of various parameters and
245 assumptions. Models to assess the impacts of biomass for BECCS on GHG concentrations
246 and climate change require details of land use change impacts, including long term nutrient
247 and productivity changes, supply chain emissions from biomass harvesting, processing, and
248 transport, combustion efficiencies of and related emissions of different fuels, and changes in
249 albedo and other biophysical processes that alter how GHGs affect the climate.

250

251 **7. Conclusion**

252 The International Energy Agency predicts that bioenergy could become the most important
253 renewable energy by 2030 (IEA, 2016). However, this depends on the implementation of
254 renewable energy strategies of key countries for the near-term deployment of BECCS at a
255 scale to meet significant energy demands. The production of biomass feedstocks are directly
256 linked to communities, farms, forests, and ecosystems from which resources are extracted.
257 Therefore, bioenergy production for BECCS has potential for significant social and justice
258 implications which could severely impede the deployment of BECCS at scale. This conflicts
259 with the suggestion that BECCS may be an early-use NET, which allows time for the
260 development of other more technically challenging NETs.

261 It is likely that NETs deployment at the huge scales envisaged in many scenarios could
262 greatly exceed our collective ability to manage carbon cycle flows, thereby risking doing
263 more harm than good. The viability of BECCS as a NET option depends on the choices made
264 throughout the supply chain. Land competition for food production and greenhouse gas
265 emissions associated with biomass cultivation, harvesting, and processing cast doubt on the
266 ability of BECCS to result in net removal of CO₂ from the atmosphere. Assessment of water,

267 land and carbon intensity of biomass supply chain and conversion technology is vital before
268 supporting the large scale deployment of BECCS. Furthermore, the impacts of large scale
269 BECCS on terrestrial biodiversity have received little consideration so far (Heck et al., 2018).
270 All current evidence shows that the anticipated high variability in the effectiveness of BECCS
271 CO₂ removal illustrates the need for a case by case analysis. The policy implication is that
272 regulating and attributing value to these systems will have to be integrated to regional
273 specificity, but how this might be achieved remains an open question (Reid, Ali & Field,
274 2019).

275 In conclusion, all available evidence points to a high variability in the possible outcomes of a
276 BECCS project, both in terms of cumulative net carbon removal over the facility's lifetime,
277 and also the time required for a given facility to start removing CO₂ from the atmosphere.
278 Furthermore, significant risks exist of perverse outcomes where the net effect is to increase
279 emissions (Fridahl & Lehtveer, 2018; Heck et al., 2018). Key factors which favour improved
280 carbon efficiencies for biomass production are: limiting the impacts of direct and indirect
281 land use changes, using carbon neutral power and organic fertilizers, prioritizing sea and rail
282 over road transport, increasing the use of carbon negative fuels, and exploiting alternative
283 biomass processing options; e.g. natural drying. However, the prospects of these achieving
284 the anticipated negative emissions by BECCS in the near term are very uncertain and
285 support the view that BECCS are not currently fit for purpose (Fuss et al., 2018).

286
287

288 **Acknowledgements:**

289 MBJ received funding from The Environmental Protection Agency, Ireland, Research
290 Programme 2014-2020, Grant No. 2016-CCRP-MS.36

291
292

293 **References:**

294 Albanito, F., Hastings, A., Fitton, N., Richards, M., Martin, M., et al. (2019). Mitigation
295 potential and environmental impact of centralized versus distributed BECCS with domestic
296 biomass production in Great Britain. *GCB Bioenergy*, 11, 1234-1252.

297 Albanito, F., Beringer, T., Corstanje, R., Poulter, B., Stephenson, A., Zawadzka, J. & Smith., P.
298 (2016). Carbon implications of converting cropland to bioenergy crops or forest for climate
299 mitigation: A global assessment. *Global Change Biology and Bioenergy*, 8, 81–95.

300

301 Alexandratos, N., & Bruinsma, J. (2012). World agriculture towards 2030/2050: The
302 2012 revision. ESA Working paper No. 12-03. Rome, FAO.

303

304 Anderson, K., & Peters, G. (2016). The trouble with negative emissions. *Science*, 354, 182-
305 183.

306
307 Baik, E., Sanchez, D.L., Turner, P. A., Mach, K.J., Field, C.B., & Benson, S. M. (2018).
308 Geospatial analysis of near-term potential for carbon-negative bioenergy in the United
309 States. *PNAS*, 115, 3290-3295.
310
311 Baldocchi, D., & Penuelas, J. (2018). The physics and ecology of mining carbon dioxide from
312 the atmosphere by ecosystems. *Global Change Biology*, 25, 1191-1197.
313
314 Batidzirai, B. Smeets, E., & Faaij, A. (2012). Harmonising bioenergy resource potentials —
315 Methodological lessons from review of state of the art bioenergy potential assessments.
316 *Renewable and Sustainable Energy Reviews*, 16, 6598 – 6630.
317
318 Bentham, M., Mallows, T., Lowndes, J., & Green, A. (2014). CO₂ storage evaluation database
319 (CO₂ Stored): The UK's online storage atlas. *Energy Procedia*, 63, 5103-13.
320
321 Bentsen, N.S., (2017). Carbon debt and payback time – Lost in the forest? *Renewable and*
322 *Sustainable Energy Reviews*, 73, 1211-1217.
323
324 Beringer, T., Lucht, W., & Schaphoff, S. (2011). Bioenergy production potential of global
325 biomass plantations under environmental and agricultural constraints. *GCB Bioenergy*, 3,
326 299-312.
327
328 Berndes, G., Hoogwijk, M., & van den Broek, R. (2003). The contribution of biomass in the
329 future global energy supply: a review of 17 studies. *Biomass and Bioenergy*, 25, 1-28.
330
331 Boysen, L. R., Lucht & W, Gerten, D. (2017). Trade-offs for food production, nature
332 conservation and climate limit the terrestrial carbon dioxide removal potential. *Global*
333 *Change Biology*, 23, 4303-4317.
334
335 Brandao, M., Kirschbaum, U. F., Cowie, A. L., & Vedel Hjuler, S. (2018). Quantifying the
336 climate change effects of bioenergy systems: comparisons of 15 impact assessment
337 methods. *GCB Bioenergy*, 11, 727-743.
338
339 Bui, M., Adjiman, C. S., Bardow, A., Anthony, E. J. et al., (2018). Carbon capture and storage
340 (CCS): the way forward. *Energy and Environmental Science*, 11, 1062-1176.
341
342 Dahmen, N., Lewandowski, I., Zibek, S., & Weidtmann, A. (2019). Integrated lignocellulosic
343 value chains in a growing bioeconomy: Status quo and perspectives. *GCB Bioenergy*, 11,
344 107-117.
345

346 Dornburg, V., van Vuuren, D., van de Ven, G., Langeveld, H., Meeusen, M. et al. (2010).
347 Bioenergy revisited: key factors in global potentials of energy. *Energy and Environmental*
348 *Science*, 3, 258-267.

349

350 EASAC. (2018). Negative emission technologies: what role in meeting Paris agreement
351 targets? *Policy Report 35*. www.easac.eu

352

353 Energy Technologies Institute (ETI). (2016). The evidence for deploying bioenergy with CCS
354 (BECCS) in the UK. Retrieved from
355 [https://d2umxnkyjne36n.cloudfront.net/insightReports/The-Evidence-for-Deploying-](https://d2umxnkyjne36n.cloudfront.net/insightReports/The-Evidence-for-Deploying-Bioenergy-with-CCS-in-the-UK.pdf?mtime=20161107110603)
356 [Bioenergy-with-CCS-in-the-UK.pdf?mtime=20161107110603](https://d2umxnkyjne36n.cloudfront.net/insightReports/The-Evidence-for-Deploying-Bioenergy-with-CCS-in-the-UK.pdf?mtime=20161107110603)

357

358 Food and Agriculture Organization. (2001). Global Forest resources assessment 2000. FAO
359 Forestry Paper 140 (p. 479). Rome, Italy.

360

361 Fajardy, M., & Mac Dowell, N. (2017). Can BECCS deliver sustainable and resource efficient
362 negative emission? *Energy Environ. Sci.* 10, 1389-1426.

363

364 Field, C.B., Campbell, J. E., & Lobell, D.B. (2007). Biomass energy: the scale of the potential
365 resource. *Trends in Ecology and Evolution*, 23, 65-72.

366

367 Fridahl, M., & Lehtveer, M. (2018) Bioenergy with carbon capture and storage (BECCS):
368 Global potential, investment preferences, and deployment barriers. *Energy Research &*
369 *Social Science*, 42, 155-166.

370

371 Fuhrman, J., McJeon, H., Doney, S. C., Shobe, W., & Clarens, A. F. (2019) From zero to hero?;
372 Why integrated assessment modelling of negative emissions technologies is hard and how
373 we can do better. *Frontiers in Climate*, 1, Article 11 doi:10.3389/fclim.2019.00011

374

375 Fuss, S., Canadell, J. G., Peters, G. P., Tavoni, M., Andrew, R. M. et al. (2014). Betting on
376 negative emissions. *Nature Climate Change*, 4, 850-853.

377

378 Fuss, S., Lamb, W. F., Callaghan, M. W., Hilaire, J., Creutzig, F. et al. (2018). Negative
379 emissions – Part 2: Costs, potentials and side effects. *Environmental Research Letters*, 13,
380 063002.

381

382 Gibbs, H. K., Ruesch, A. S., Achard, F., Clayton, M. K., Holmgren, P., Ramankutty, N., et al.
383 (2010). Tropical forests were the primary sources of new agricultural land in the 1980s and
384 1990s. *Proceedings of the National Academy of Sciences*, 107(38), 16732-16737.

385

386 Gough, C., Garcia-Freites, S., Jones C, & Manders, S. (2018). Challenges to the use of BECCS
387 as a keystone technology in pursuit of 1.5°C. *Global Sustainability*, 1, e5, 1-9.

388

389 Haberl, H., Beringer, T., Bhattacharya, S. C., Erb, K-H., & Hoogwijk, M. (2010). The global
390 technical potential of bio-energy in 2050 considering sustainability constraints. *Current*
391 *Opinion in Environmental Sustainability*, 2, 394-403.

392

393 Hastings, A. (2017). Bioenergies impacts on Natural Capital and Ecosystem Services: A
394 comparison of biomass and coal fuels. In. Z. Qin, U. K. Mishra, & A. Hastings (Eds.) *Bioenergy*
395 *and Land Use Change* (pp. 83-97) AGU/Wiley.

396

397 Heck, V., Gerten, D., Lucht, W., & Popp, A. (2018) Biomass-based negative emissions difficult
398 to reconcile with planetary boundaries. *Nature, Climate Change* doi.org/10.1038/s41558-
399 017-0064-y

400

401 IEA (2016). *Energy Technology Perspectives 2016: Towards Sustainable Urban Energy*
402 *Systems*. Paris. OECD/IEA

403

404 IPCC (2019). *Climate change and land. An IPCC special report on climate change,*
405 *desertification, land degradation, sustainable land management, food security, and*
406 *greenhouse gas fluxes in terrestrial ecosystems.* P.R. Shukla, J. Skea, E. Calvo Buendia, V.
407 Masson-Delmotte, H.-O. Pörtner, D. C. Roberts, P. Zhai, R. Slade, S. Connors, R. van Diemen,
408 M. Ferrat, E. Haughey, S. Luz, S. Neogi, M. Pathak, J. Petzold, J. Portugal Pereira, P. Vyas, E.
409 Huntley, K. Kissick, M, Belkacemi, J. Malley, (eds.). In press.

410

411 Kato, E., & Yamagata, Y. (2014) BECCS capability of dedicated bioenergy crops under a
412 future land-use scenario targeting net negative carbon emissions. *Earth's Future*, 2, 421-439.

413

414 Kemper, J. (2015). Biomass and carbon dioxide capture and storage: a review. *International*
415 *Journal of Greenhouse Gas Control*, 40, 401-430.

416

417 Lewandowski, I. (2015). Securing a sustainable biomass supply in a growing bioeconomy.
418 *Global Food Security*, 6, 34-42. doi: 10.1016/j.gfs.2015.10.001.

419

420 Mason, P.M., Glover, K., Smith, J.A.C., Willis, K.J., Woods, J., & Thompson, I.P. (2015). The
421 potential of CAM crops as a globally significant bioenergy resource: moving from 'fuel or
422 food' to 'fuel and more food'. *Energy & Environmental Science*, 8, 2320-2329.

423

424 McKechnie, J., Colombo, S., Chen, J., Mabee, W., & Maclean, H. (2011). Forest bioenergy or
425 forest carbon? Assessing trade-offs in greenhouse gas mitigation and wood-based fuels.
426 *Environmental Science and Technology*, 45, 789-795.

427

428 Madsen, K., & Bentsen, N. S. (2018). Carbon debt payback time for a biomass fired CHP plant
429 – A case study from Northern Europe. *Energies*, 11, 807; doi:10.3390/en11040807.

430
431 Melillo, J. M., Reilly, J. M., Kicklighter, D. W., Gurgel, A. C., Cronin, T. W. *et al.* (2009).
432 Indirect emissions from biofuels: how important? *Science*, 326, 1397-1399.
433
434 Minx, J. C., Lamb, W., Callaghan, M., Bornmann, L., Fuss, S. *et al.* (2017). Fast growing
435 research on negative emissions. *Environmental Research Letters*, 12, 035007.
436
437 Muri, H. (2018). The role of large-scale BECCS in the pursuit of the 1.5°C target: an earth
438 system model perspective. *Environmental Research Letters*, 13, 044010.
439
440 National Academies of Sciences (NAS) (2019). Negative Emissions Technologies and Reliable
441 Sequestration: A Research Agenda. Washington, DC: The National Academies Press. doi:
442 <https://doi.org/10.17226/25259>.
443
444 Nemet, G. F., Callaghan, M. W., Creutzig, F., Fuss, S., Hartmann, J. *et al.* (2018). Negative
445 emissions –Part 3: Innovation and upscaling. *Environmental Research Letters*, 13, 063003.
446
447 Odeh, N. A. and Cockerill, T.T. (2008). Life cycle GHG assessment of fossil fuel power plants
448 with carbon capture and storage. *Energy Policy* 36, 367-80.
449
450 Overmars, K., Edwards, R., Padella, M., Prins, A. G., & Marelli, L. (2015). Estimates of indirect
451 land use change from biofuels based on historical data. Luxembourg: Publications Office of
452 the European Union.
453
454 Reid, W. V., Ali, M. K., & Field, C. B. (2019). The future for bioenergy. *Global Change Biology*
455 Doi: 10.1111/GCB.14883.
456
457 Röder, M., Whittaker, C., & Thornley, P. (2015). How certain are greenhouse gas reductions
458 from bioenergy? Life cycle assessment and uncertainty analysis of wood pellet-to-electricity
459 supply chains from forest residues. *Biomass and Bioenergy*, 79, 50– 63.
460
461
462 Roder, M., & Thornley, P. (2016). Bioenergy as climate change mitigation option within a 2°C
463 target – uncertainties and temporal challenges of bioenergy systems. *Energy, Sustainability*
464 *and Society*, 6, doi: 10.1186/s13705-016-0070-3.
465
466 Sanchez, L., Johnson, N., McCoy, S. T., Turner, P.A., Mach, K. J. (2018) Near-term
467 deployment of carbon capture and sequestration from biorefineries in the United States.
468 *PNAS*, 115, 4875-4880.
469
470 Searle, S. Y., & Malins, C. J. (2014). Will energy crop yields meet expectations? *Biomass &*
471 *Bioenergy*, 65, 3-12.

472
473 Searle, S., & Malins C. (2015). A reassessment of global bioenergy potential in 2050. *GCB*
474 *Bioenergy*, 7, 328-336.
475
476 Smith, L. J., & Torn, M. S., (2013). Ecological limits to terrestrial biological carbon dioxide
477 removal. *Climate Change*, 118, 89-103.
478
479 Smith, P., Davis, S. J., Creutzig, F., Fuss, S., Minx, J. et al. (2015). Biophysical and economic
480 limits to negative CO₂ emissions. *Nature Climate Change*, 6, 42-50.
481
482 Smith, P., Haszeldine, R. S., Smith, S. M. (2016). Preliminary assessment of the potential for,
483 and limitations to, terrestrial negative emission technologies in the UK. *Environmental*
484 *Science: Processes & Impacts*, 18, 1400–1405
485
486 Smith, P., Bustamante, M., Ahammad, H., Clark, H., Dong, H. et al. (2014). Agriculture,
487 Forestry and Other Land Use (AFOLU). In: *Climate Change 2014: Mitigation of Climate*
488 *Change. Contribution of Working Group III to the Fifth Assessment Report of the*
489 *Intergovernmental Panel on Climate Change.* Edenhofer, O., R. Pichs-Madruga, Y. Sokona, E.
490 Farahani, S. Kadner, K. Seyboth, A. Adler, I. Baum, S. Brunner, P. Eickemeier, B. Kriemann, J.
491 Savolainen, S. Schlömer, C. von Stechow, T. Zwickel and J.C. Minx (eds.). Cambridge
492 University Press, Cambridge, United Kingdom and New York, NY, USA.
493
494 Stermann, J. D., Siegel, L., & Rooney-Varga, N. (2018). Does replacing coal with wood lower
495 CO₂ emissions? Dynamic lifecycle analysis of wood bioenergy. *Environmental Research*
496 *Letters*, 13, 015007
497
498 Tanzer, S.E., & Ramirez, A. (2019). When are negative emissions negative emissions? *Energy*
499 *and Environmental Science*, 12, 1210-1218.
500
501 Woods, J., Lynd, L. R., Laser, M., Batistella, M., Victoria, D. C., Kline, K., & Faaij, A. (2015).
502 Land and bioenergy. In: *Bioenergy & Sustainability: Bridging the Gaps 72.* SCOPE, Paris.
503 France, 258–301, (ISBN 978-2-9545557-0-6).
504
505 Valin, H., Peters, D., van den Berg, M., Frank, S., Havlík, P., Forsell, N., & Hamelinck, C.
506 (2015). The land use change impact of biofuels consumed in the EU Quantification of area
507 and greenhouse gas impacts, ECOFYS, Netherlands B.V. www.ecofys.com
508
509 Vaughan, N. E., & Gough, C. (2016). Expert assessment concludes negative emissions
510 scenarios may not deliver. *Environmental Research Letters*, 11(9), 095003
511

512 Vaughan, N. E., Gough, C., Mander, S., Littleton, E. W., Welfle, A. Gernaat, H. J., van Vuuren
 513 D. P. (2018) Evaluating the use of biomass energy with carbon capture and storage in low
 514 emission scenarios. *Environmental Research Letters*, 13, 044014

515
 516
 517
 518
 519
 520

521 Table 1. Published estimates of global bioenergy technical production potential (*i.e.* amount
 522 of biomass energy that can be supplied globally given current expectations on technology,
 523 food demand and environmental constraints), and potential indirect land use change (iLUC)
 524 GHG emissions from the demand for food crops in new areas (iLUC) due to the conversion
 525 of current croplands to biomass feedstock production.

526

<i>Reference</i>	<i>Bioenergy Production Potential</i>	<i>Comments</i>
Haberl et al. (2010)	160-270 EJ/yr in 2050	'Scientific studies required in order to be more precise'
Dornburg et al. (2010)	200-500 EJ/yr in 2050	
Berndes et al. (2003)	100->400 EJ/yr in 2050	Review of 17 studies
Beringer et al. (2011)	130-270 EJ/yr in 2050	Used LPJmL DGVM model
Rogier et al. (2012)	793 EJ/yr currently	
Kemper et al. (2015)	50->1000 EJ/yr currently	'most likely range'
Fuss et al. (2018)	60-1548 EJ/yr in 2050	
<i>Reference</i>	<i>GHG emissions from iLUC</i>	<i>Production system</i>
Valin et al. (2015)	17 gCO ₂ -eq/MJ	Forest residues
Overmars et al. (2015)	2-3 gCO ₂ -eq/MJ	Cereal straw & stover
Valin et al. (2015)	0-16 gCO ₂ -eq/MJ	Cereal straw & stover
Valin et al. (2015)	-12 gCO ₂ -eq/MJ	Swichgrass & miscanthus
Melillo et al. (2009)	275-285 gCO ₂ -eq/MJ	Eucalyptus, swichgrass & poplar

527
 528
 529

530 Figure Legend:

531

532 Figure 1. The simplistic vision of BECCS is that one ton of CO₂ captured in the growth of
 533 biomass would equate to one ton of CO₂ sequestered geologically, which we can regard as a
 534 carbon efficiency of 1 (*i.e.* Gate to Gate with carbon neutrality). This simplistic concept of
 535 carbon neutrality in the bioenergy debate, however, is far from the reality. Depending on
 536 the different technology assessments boundaries applied to the BECCS scenario (*e.g.* Cradle
 537 to Grave), GHG emissions are emitted throughout the biomass supply-chain reducing the

538 carbon efficiency of BECCS to less than 50% (EASAC, 2018). In bioenergy systems, indirect
539 land use change (iLUC) also needs to be included to achieve a full picture of the system
540 impacts (*i.e.* Cradle to Grave with iLUC).

541

542